HOW DO PLANTS RESPOND TO PATCH AREA AND ITS DISTRIBUTION PATTERN IN HORQIN SAND LAND, CHINA

WU, $J^{1,2}_{\cdot}$ – HOU, X. Z.¹ – XU, C. L.³ – LIU, Z. M.^{2*}

¹Liaoning Ecological Engineering Vocational College, Shenyang 110001, P R China

²Institute of Applied Ecology, Chinese Academy of Sciences, Shenyang 110016, P R China

³Experimental Forest Farm of Liaoning province, Fushun 113300, P R China

*Corresponding author e-mail: 0227wujing@163.com

(Received 30th Sep 2018; accepted 29th Nov 2018)

Abstract. To elucidate the plant response to habitat fragmentation, 18 interdune lowlands with different sizes in active sand dunes of Horqin Sand Land were selected, and the interdune lowland was considered as fragmented habitat patch. In our study, the effect of patch size on plant distribution pattern was explored and different protocols of species diversity conservation were proposed. Our results showed different plant sensitivities to habitat fragmentation: type I, species are restricted by patch area and distributed regularly in fragments; type II, species are not restricted by patch area, and their distribution is irregular in fragments; type II, species mainly distributed in large fragments; type IV, species mainly distributed in middle-sized fragments. Exploring the effects of fragmentation habitat size on plant species diversity and its distribution pattern will provide theoretical basis for plant diversity conservation in semi-arid sand dunes. **Keywords:** *biodiversity, habitat loss, habitat isolation, patch, landscape*

Introduction

Habitat fragmentation has been considered as one of the major threats to biodiversity (Brunet et al., 2011; Wu et al., 2013; Murphy and Romanuk 2014; Matthews et al., 2014; Ducatez and Shine, 2017). Fragmentation processes often result in "patchy" habitat segments, called patchy habitat. As the degree of fragmentation increases, the original patch is isolated from the highly altered retrogressive landscape and gradually recedes and eventually develops into a biogeographic "habitat islands", producing a range of ecological or biological effects at different levels of population, community, ecosystem and even landscape.

The size, environmental heterogeneity and marginal effects of habitat patches have important effects on species richness and abundance in patches (Benedick et al., 2006; Santos et al., 2008; Leal et al., 2012; Fahrig et al., 2015). Different species have different sensitivity, adaptability and tolerance to landscape fragmentation (Haila, 2002; Hill et al., 2003; Swihart et al., 2003; Kolb et al., 2005; Hudson et al., 2017). The influence of habitat patches such as Qiandao Lake and Three Gorges Reservoir on plant diversity has been reported in China (Ding, 2005). Many studies have focused on the real islands in geography (Halley et al., 2014; Phillips et al., 2018), but less attention has been paid to the ecological effects of habitat patches in the broad sense of biogeography and ecology. As a relatively independent vegetation unit in sand dune ecosystem, the interdune lowlands have unique patch properties in sand dune landscape, which can be regarded as "fragmented habitat islands" in sand dune ecosystem, and play an important role in determining plant diversity and distribution pattern of sand dune ecosystem. However, the impact of fragmented patch area in the sand dune ecosystems of semi-arid regions (i.e., interdune lowlands) on plant diversity and island accumulation has not been reported.

Therefore, we investigated the plant diversity of interdune lowlands in the active dune ecosystem of the Horqin Sandy land, and conducted regression analysis on the relationship between the relative frequency and abundance of typical species and fragmented island area. This study is aiming to explore the plant diversity protection scheme applicable to different ecological groups and reveal the processes and mechanisms of the impacts of the habitat island on the plant diversity. This study can provide a theoretical basis for biodiversity protection and guides the practice of biodiversity protection in semi-arid areas.

Material and Methods

Study site

The study site is located at Wulanaodu region $(42^{\circ}47' - 43^{\circ}25' \text{ N}, 118^{\circ}38' - 120^{\circ}43' \text{ E}, 480 \text{ m a.s.l.})$, eastern Inner Mongolia, China (*Figure 1*). The region has a semi-arid climate. The annual average temperature is 6.3° C. The annual average rainfall is ca. 340 mm, most of which is received during June to September. The windy season is from March to May. The growing season starts in late April and ends in late September.



Figure 1. Location map and five sampling sites in Horqin sandy land

The active sand dune areas, 15–35 m in height, are advancing at a speed of ca. 6–7 m year-1, and, following the aforementioned fencing since the 1980s, are permanently closed off from the natural grassland (>600 ha). The natural grassland and the active sand dunes experience the same climatic conditions, i.e. similar rainfall, temperature, humidity etc. The vegetation before the intense grazing period was dominated by perennial grasses, such as *Phragmites communis* and *Calamagrostis epigeios*, whereas after the onset of grazing pasammophilous species, such as *Artemisia wudanica* and *Agriophyllum squarrosum*, became more established.

Experimental procedures

We selected 18 interdune lowlands ranging in size from 0.10 to 5.64 hm^2 in 2012 (*Table 1*) (Liu et al., 2003). We used grid sampling at a resolution of 10×10 m. We established transects divided into 10-m intervals in each plot, all running in the vertical direction of the prevailing wind, and then selected 1×1 m quadrats, 10-m-spaced, along each transect (Wu et al., 2016). The species richness and abundance in the sample were recorded. For bunchgrasses (e.g. *C. squarrosa*), we counted the number of clusters to obtain the abundance, whereas for clonal species (e.g. *S. gordejevii* and *P. communis*) we counted the number of ramets, and for discrete species (e.g. *A. squarrosum*) the number of individuals were counted. The frequency of each species was determined within each plot.

Plot number	Geographical coordinates of the center of the interdune lowland	Size (hm ²)	Dam Height (m)	
1	42°59.900 N, 119°38.009 E	0.05	25	
2	42°59.367 N, 119°38.789 E	0.27	28.5	
3	42°58.089 N, 119°38.169 E	0.30	25.5	
4	42°58.170 N, 119°37.886 E	0.32	25	
5	43°00.964 N, 119°38.887 E	0.60	24	
6	42°59.789 N, 119°37.236 E	0.66	25	
7	42°59.767 N, 119°37.709 E	0.67	22.5	
8	42°59.640 N, 119°37.886 E	1.00	25	
9	42°59.805 N, 119°37.964 E	1.35	24.5	
10	43°00.001 N, 119°37.870 E	2.27	23	
11	42°59.862 N, 119°37.974 E	2.94	19	
12	42°59.958 N, 119°37.844 E	3.22	23	
13	42°59.442 N, 119°38.749 E	4.35	26.5	
14	43°00.336 N, 119°37.609 E	5.01	25	
15	42°00.498 N, 119°37.154 E	5.25	24.5	
16	42°00.598 N, 119°38.504 E	6.19	26	
17	42°00.156 N, 119°38.454 E	6.84	22	
18	42°00.453 N, 119°37.231 E	11.3	24.5	

Table 1. Basic information of 18 selected interdune lowlands

Species diversity measure

- Species abundance: number of species within a community.
- Frequency of a species (%) =occurrence of the certain species/number of total samples.
- Abundance of a species (%) = total plants of certain species/ total plants of total species.

Classification of typical plant responses to habitat fragmentation

Based on the differences of typical plant responses to the habitat area, plants were divided into the five categories (*Table 2*).

Data analysis

All the statistic analysis was conducted in SPSS software. Model selection was based on maximum R^2 , and residual analysis showed that the model fit was adequate.

Statistical significance was determined at P = 0.05. We analyzed the relationships between patch area and frequency and relative abundance for 6 type's species.

Туре	Relationship	Distribution
Ι	Significant	Regular
II	No significant	Irregular
III	Significant	Mainly distributed in large fragments
IV	No significant	Mainly distributed in small fragments
V	No significant	Mainly distributed in middle-sized fragments

Table 2. Classification of typical plant responses to habitat fragmentation

Results

Species composition

According to the data of 18 interdune lowlands, there are 114 species belonging to 36 families in the patch habitat of horqin Sandy land. Among them, 25 species belong to *Compositae*, 13 species belong to *Graminae*, 11 species belong to *Leguminosae*, 6 species belong to *Cyperaceae*, 7 species belong to *Chenopodiaceae*, and the remaining 57 species belong to 31 families. Among them, 10 species is Psammophytes, 76 species are meadows plants and 22 species are steppes (*Table 3*).

Table 3. Species	list in the	18 selected	patchy habitats
------------------	-------------	-------------	-----------------

Number	species	General	Ecological group	Life form	Frequency
1	Artemisia halodendron	Compositae	Р	SS	0.99
2	Artemisia wudanica		Р	SS	1.36
3	Artemisia frigida		Р	SS	1.24
4	Inula salsoloides		Р	PH	10.7
5	Artemisia gmelinii		LMS	SS	0.33
6	Artemisia lavandulaefolia		LMS	PH	23.5
7	Erigeron acer		LMS	BH	0.06
8	Eupatorium lindleyanum		LMS	PH	0.07
9	Hypochoeris grandiflora		LMS	PH	0.03
10	Inula britannica		LMS	PH	8.78
11	Lactuca indica		LMS	BH	1.06
12	Lactuca tatarica		LMS	PH	5.70
13	Leibnitzia anandria		LMS	PH	6.89
14	Taraxacum mongolicum		LMS	PH	16.7
15	Taraxacum borealisinense		LMS	PH	3.45
16	Carduus nutans		LMS	BH	0.02
17	Cirsium segetum		LMS	PH	0.46
18	Ixeris chinensis		LMS	PH	34.1
19	Scorzonera capito		LMS	PH	0.05
20	Senecio jacobacea		LMS	PH	0.42
21	Sonchus brachyotus		LMS	PH	10.4
22	Serratula cardunculus		LMS	PH	0.04
23	Xanthium sibiricum		LMS	AH	0.07

Number	species	General	Ecological group	Life form	Frequency
24	Heteropappusaltaicus		STS	PH	0.18
25	Agriophyllum squarrosum	Chenopodiaceae	e P	AH	0.34
26	Corispermum candelabrum	-	Р	AH	5.77
27	Salix gordejevii	Salicaceae	Р	S	34.5
28	Caragana microphylla	Leguminosae	Р	S	1.56
29	Hedysarum fruticosum	-	Р	SS	3.76
30	Agrostis clavata	Gramineae	LMS	PH	16.8
31	Arthraxon hispidus		LMS	AH	9.87
32	Calamagrostis epigeios		LMS	PH	46.6
33	Pennisetum alopecuroides		LMS	PH	1.31
34	Phragmites communis		LMS	PH	60.4
35	Miscanthussacchariflorus		LMS	PH	15.6
36	Echinochloa frumentacea		LMS	PH	0.31
37	Septoria mougeotii		LMS	PH	0.02
38	Chloris virgata		LMS	AH	10.8
39	Eragrostis pilosa		LMS	AH	30.2
40	Setaria viridis		LMS	AH	0.98
41	Pennisetum centrasiaticum		LMS	PH	1.67
42	Astragalus adsurgens	Leguminosae	LMS	PH	0.45
43	Glycine soja		LMS	AH	3.06
44	Kummerowia striata		LMS	PH	2.64
45	Melilotus suaveolens		LMS	ABH	0.67
46	Trigonella korshinskyi		LMS	AH	0.01
47	Vicia amoena		LMS	PH	0.34
48	Swainsonia salsula		LMS	SS	1.78
49	Thermopsis lanceolata		LMS	PH	1.83
50	Radix Glycyrrhizae		STS	PH	0.65
51	Bolboschoenus compacts	Cyperaceae	LMS	PH	5.91
52	Bolboschoenus planiculmis		LMS	PH	0.37
53	Carex caespitosa		LMS	PH	10.6
54	Carex duriuscula		LMS	PH	12.3
55	Heleochalis intersita		LMS	PH	0.26
56	Scirpus tabernaemontani		LMS	PH	0.02
57	Populus spp.	Salicaceae	LMS	S	5.76
58	Salix microstachya		LMS	S	31.5
59	Salix mongolica		LMS	S	0.87
60	Alisma orientale	Alismataceae	LMS	PH	0.04
61	Sagittaria trifolia		LMS	PH	0.05
62	Chenopodium acuminatum	Chenopodiaceae	LMS	AH	0.04
63	Chenopodium glaucum		LMS	AH	3.67
64	Bassia dasyphylla		STS	AH	0.34
65	Chenopodium aristatum		STS	AH	0.22
66	Salsola ruthenica		STS	AH	0.12
67	Polygonum hydropiper	Polyonaceae	LMS	PH	0.02
68	Polygonum lapathifolium		LMS	AH	0.01
69	Polygonum thunbergu		LMS	AH	0.01

APPLIED ECOLOGY AND ENVIRONMENTAL RESEARCH 17(1):1223-1234. http://www.aloki.hu • ISSN 1589 1623 (Print) • ISSN 1785 0037 (Online) DOI: http://dx.doi.org/10.15666/aeer/1701_12231234 © 2019, ALÖKI Kft., Budapest, Hungary

Number	species	General	Ecological group	Life form	Frequency
70	Polygonum laxmanni		P	PH	0.01
71	Equisetum ramosissimum	Equisetaceae	LMS	PH	0.02
72	Eguisetum sylvalicum	•	LMS	PH	0.02
73	Euphorbia humifusa	Euphorbiaceae	LMS	AH	0.02
74	Gentiana squarrosa	Gentianaceae	LMS	PH	0.02
75	Plantago depressa	Plantaginaceae	LMS	PH	3.09
76	Glaux maritima	Primulaceae	LMS	PH	0.05
77	Halerpestes cymbalaria	Ranumculaceae	LMS	PH	5.93
78	Potentilla discolor	Rosaceae	LMS	PH	1.32
79	Potentillae Chinensis		LMS	PH	1.02
80	Potentilla supina		LMS	PH	0.67
81	Potentilla anserina		LMS	PH	0.66
82	Typha minima	Typhaceae	LMS	PH	1.33
83	Artemisia laciniata	Compositae	STS	PH	1.78
84	Artemisia scoparia		STS	ABH	3.65
85	Artemisia sieversiana		STS	ABH	1.34
86	Heteropappus altaicus		STS	PH	0.34
87	Lespedeza davurica	Leguminosae	STS	SS	10.9
88	Oxytropis ramosissima		STS	PH	0.02
89	Sophora flavescens		STS	PH	1.98
90	Cynanchum sibiricum	Asclepiadaceae	STS	PH	0.31
91	C. chinense		STS	PH	0.01
92	Cleistogenes squarrosa	Gramineae	STS	PH	6.81
93	Galium verum	Rubiaceae	STS	PH	0.28
94	Rubiaschumahhiaha		Р	PH	0.01
95	Ulmus pumila	Ulmaceae	STS	S	0.05
96	Linaria vulgaris	Scrophulariaceae	LMS	PH	0.01
97	Portulaca oleracea	Portulacaceae	STS	AH	0.05
98	Spiked Loosestrlfe	Lythraceae	LMS	PH	0.01
99	Lythrum virgatum		LMS	PH	0.01
100	Erodium stephanianum	Geraniaceae	LMS	PH	0.01
101	Allium odorum	Liliaceae	LMS	PH	0.01
102	Asparagus brachyphyllus		LMS	PH	0.01
103	Asparagua dahuricus		LMS	PH	0.01
104	Viola philippica	Violaceae	LMS	PH	0.01
105	Viola prionantha		LMS	PH	0.01
106	Tragus bertesonianus	Rutaceae	STS	PH	0.01
107	Lappula echinata	Boraginaceae	STS	ABH	0.01
108	Radix Arnebiae		STS	PH	0.01
109	Scutellaria baicalensis	Labiatae	STS	PH	0.01
110	Tirbulus terrestris	Zygophyllaceae	LMS	AH	0.01
111	Saposhnikovia divaricata	Umbelliferae	STS	PH	0.01
112	Silene jenisseensis	Caryophyllaceae	STS	PH	0.02
113	Cuscuta chinensis	Convolvulaceae	STS	AH	0.01
114	Betula platyphylla	Betula	LMS	S	0.01

The responses of typical plants to fragmented patch area

Type I (species is restricted by patch area and distributed regularly in fragments)

The frequency and relative abundance of *Artemisia wudanica*, *Phragmites communis* decreased logarithmically. The frequency and relative abundance of *Lespedeza davurica* showed a change of binomial function with the increase of the area of the fragmented habitat, when the area of plaque is around 2 hm², its frequency is lowest, and its relative abundance tends to increase with patch area (*Figure 2*).



Figure 2. Changes in the frequency and abundance of Artemisia wudanica, Phragmites communis, Lespedeza davurica with increasing interdune lowland size

Type II (species is not restricted by patch area, and their distribution is irregular in fragments)

There was no significant change in the frequency and relative abundance of *Salix gordejevii* with the increase of plaque habitat area. *Salix gordejevii* appeared in different interdune lowlands, and the relative abundance of *Salix gordejevii* decreased with the increase of patch area (*Figure 3*).

Type III (species mainly distributed in large fragments)

The frequency and relative abundance of *Eragrostis pilosai* increased significantly with the increase of patch area, that is, the population size and distribution of *Eragrostis pilosai* declined sharply (*Figure 4*).

Type IV (species mainly distributed in small fragments)

The frequency and relative abundance of *Vicia amoena* showed no significant change with the increase of patch area. *Vicia amoena* is more frequent in small and medium-sized interdune lowlands; the population size of *Vicia amoena* varies little with increased habitat fragmentation (*Figure 5*).



Figure 3. Changes in the frequency and relative abundance of Salix gordejevii population with interdune lowland area

Figure 4. Changes in the frequency and relative abundance of Eragrostis pilosai population with interdune lowland area



Figure 5. Changes in the frequency and relative abundance of Vicia amoena population with interdune lowland area

APPLIED ECOLOGY AND ENVIRONMENTAL RESEARCH 17(1):1223-1234. http://www.aloki.hu • ISSN 1589 1623 (Print) • ISSN 1785 0037 (Online) DOI: http://dx.doi.org/10.15666/aeer/1701_12231234 © 2019, ALÖKI Kft., Budapest, Hungary

Type V (species mainly distributed in middle-sized fragments)

The frequency and relative abundance of *Chenopodium aristatum* showed no significant change with the increase of patch area, the frequency of *Chenopodium aristatum* was higher in medium size patch and lower in small and large size patch. The population size of *Chenopodium aristatum* had no significant change with the increase of habitat fragmentation. There was no significant change in the frequency and relative abundance of *Sophora miltiorrhiza* with the increase of patch area, and the size of population was not significant change with the increase of habitat fragmentation (*Figure 6*).



Figure 6. Change in the frequency and relative abundance of Chenopodium aristatum and Sophora miltiorrhiza with interdune lowland area

Discussion

Contrary to the existing view that species frequency show a negative trend with decreasing fragment area (Hill and Curran, 2003; Echeverría et al., 2007; Kornera and Jeltschb, 2008), our results showed that some species responded positively to habitat fragmentation and the species tolerance to habitat fragmentation was different. The plant-environmental feedback process in the dune system has a significant effect on plant growth and reproduction (Hambäck et al., 2007; Giladi et al., 2014; Winfree et al., 2015). The feedback process of *Artemisia wudanica* to wind erosion is as follows: After *Artemisia wudanica* is blown by the wind, the plant will fall down at a certain wind erosion depth, and the sand blocking area will increase accordingly. The microenvironment of the buried parts of the plants has changed. For example, 1) the soil volume has increased and the soil water content has increased; 2) mycorrhizal fungus activity capacity increased; 3) in the effective use of resources increased in the soil.

These changes have created favorable conditions for the adventitious roots on the sandy branches. In the case of a large number of adventitious roots, the growth of the plant increases, and the ability to block sand increases.

Reeds can appear in meadow grasslands and in the interdune lowlands. On the one hand, they belong to meadows, and on the other hand, they have the ability to adapt to wind sand. Previous studies have concluded that reeds can quickly invade and settle in habitats with intense sand activity due to its strong cloning and reproduction ability, (Liu et al., 2006; Liu et al., 2008). Therefore, in the broken patches of mobile dunes, especially in small areas with intense sand activity (Ma et al., 2010; Zobel et al., 2010; Wu et al., 2013), reeds show the ability to adapt to sandstorm interference (Liu et al., 2007). Reeds can be rapidly expanded from lowland to windward slopes through Rhizome extension and asexual strains. As pioneers, they create suitable habitat conditions for other species to invade, drive new vegetation succession in transition zones, and reduce wind and sand fluidity.

Lespedeza davurica has a persistent soil seed bank and can also ensure its population development under strong interference, while playing an important role in maintaining community diversity and system stability (Li et al., 2006). Lespedeza davurica population can expand the adjacent space by increasing the number of branches. The high-density population of Lespedeza davurica expands the plant elevation space in the form of lateral branches, that is, transverse space. The low-density population of Lespedeza davurica branches, the longitudinal space by increasing its branches, gradually occupied the horizontal space (Li, 2005).

Conclusions

In the practice of species diversity conservation, programmers should be tailored to the sensitivity of the species to habitat fragmentation. For example, with the increase in the area of patches, the frequency and relative diversity of *Artemisia wudanica*, *Phragmites communis*, and *Vicia amoena* are all reduced. Therefore, when protecting *Artemisia wudanica*, *Phragmites communis*, and *Vicia amoena*, the focus should be on protecting small islands. The frequency and relative abundance of *Lespedeza davurica* increase with the increase of patch area. Therefore, when protecting *Lespedeza davurica*, we should focus on a large area of islands; the frequency and relative abundance of *Chenopodium aristatum* and *Sophora flavescens* are higher in the middle area islands, therefore, when protecting *Chenopodium aristatum* and *Sophora flavescens*, the middle area islands should be protected.

Acknowledgments. We thank Jianqiang Qian and Yongming Luo for field assistance. The study was financially supported by the National Natural Science Youth Foundation of China (31600443) and the Liaoning Natural Science Foundation (201800170).

REFERENCES

[1] Benedick, S., Hill, J. K., Mustaffa, N., Chey, V. K., Maryati, M., Searle, J. B., Schilthuizen, M., Hamer, K. C. (2006): Impacts of rain forest fragmentation on butterflies in northern Borneo: Species richness, turnover and the value of small fragments. – J Appl Ecol 43: 967-977.

- [2] Brunet, J., Valtinat, K., Mayr, L. M., Felton, A., Lindbladh, M., Bruu, H. H. (2011): Understory succession in post-agricultural oak forests: habitat fragmentation affects forest specialists and generalists differently. – For. Ecol. Manage. 262: 1863-1871.
- [3] Ding, L. Z. (2005): Islanding landscape in thousand-island lake region and its effect on plant diversity. Hangzhou, Zhejiang University.
- [4] Ducatez, S., Shine, R. (2017): Drivers of extinction risk in terrestrial vertebrates. Conserv. Lett. 10: 186-194.
- [5] Echeverría, C., Newton, A. C., Lara, A., Benayas, J. M. R., Coomes, D. A. (2007): Impacts of forest fragmentation on species composition and forest structure in the temperate landscape of southern Chile. – Global Ecol. Biogeogr. 16: 426-439.
- [6] Fahrig, L., Girard, J., Duro, D., Pasher, J., Smith, A., Javorek, S., King, D. J., Lindsay, K. E., Mitchell, S. W., Tischendorf, L. (2015): Farmlands with smaller crop fields have higher within-field biodiversity. Agric. Ecosyst. Environ. 200: 219-234.
- [7] Giladi, I., May, F., Ristow, M., Jeltsch, F., Ziv, Y., Triantis, K. (2014): Scale-dependent species-area and species-isolation relationships: a review and a test study from a fragmented semi-arid agro-ecosystem. J. Biogeogr. 41: 1055-1069.
- [8] Haila, Y. (2002): A conceptual genealogy of fragmentation research: From island biogeography to landscape ecology. Ecol Appl 12: 321-334.
- [9] Halley, J. M., Sgardeli, V., Triantis, K. A. (2014): Extinction debt and the species–area relationship: a neutral perspective. Global Ecol. Biogeogr. 23: 113-123.
- [10] Hambäck, P. A., Summerville, K. S., Steffan-Dewenter, I., Krauss, J., Englund, G., Crist, T. (2007): Habitat specialization, body size, and family identity explain lepidopteran density-area relationships in a cross-continental comparison. – Proc. Natl Acad. Sci. USA. 104: 8368-8373.
- [11] Hill, J. L., Curran, P. L. (2003): Area, shape and isolation of tropical forest fragments: effects on tree species diversity and implications for conservation. – J Biogeogr 30: 1391-1403.
- [12] Hudson, L. N., Newbold, T., Contu, S., Hill, S., Lysenko, I., Palma, A. D., Phillips, H. R. P., Alhusseini, T. I., Bennett, D. J., Bugter, R. J. F., Buscardo, E. (2017): The database of the PREDICTS (Projecting Responses of Ecological Diversity in Changing Terrestrial Systems) project. Ecol. Evol. 7: 145-188.
- [13] Kolb, A., Diekmann, M. (2005): Effects of life-history traits on responses of plant species to forest fragmentation. – Conserv Biol 19: 929-938.
- [14] Kornera, K., Jeltschb, F. (2008): Detecting general plant functional type responses in fragmented landscapes using spatially-explicit simulations. Ecol. Model. 210: 287-300.
- [15] Leal, I. R., Filgueiras, B. K., Gomes, J. P., Luciana, B., Bullet, I., Andersen, A. N., Conserv, B. (2012): Effects of habitat fragmentation on ant richness and functional composition in Brazilian Atlantic forest. – Biodivers. Conserv. 21(7): 1687-1701.
- [16] Li, L. (2005): Study on germination and growth of Lespedeza davurica. Changchun. Dongbei Normal University.
- [17] Li, X. H., Li, X. L., Jiang, D. M., Liu, Z. M. (2006): Germination strategy and ecological adaptability of Eragrostis pilosa. – J Appl Ecol 17(4): 607-670. (in Chinese with English abstract).
- [18] Liu, B., Liu, Z. M., Guan, D. X. (2008): Seedling growth variation in response to sand burial in four Artemisia species from different habitats in the semi-arid dune field. – Trees-struct Funct 22: 41-47.
- [19] Liu, J. G., Ouyang, Z. Y., Pimm, S. L. (2003): Protecting China's Biodiversity. Science, 300: 1240-1241.
- [20] Liu, Z. M., Yan, Q. L., Baskin, C. C., Ma, J. L. (2006): Burial of canopy-stored seeds in the annual psammophyte Agriophyllum squarrosum (Chenopodiaceae) and its ecological significance. – Plant Soil 288: 71-80.

- [21] Liu, Z. M., Li, X. L., Yan, Q. L., Wu, J. G. (2007): Species richness and vegetation pattern in interdune lowlands of an active dune field in Inner Mongolia, China. – Biol Conserv. 140: 29-39.
- [22] Ma, J. L., Liu, Z. M., Zeng, D. H., Liu, B. (2010): Aerial seed bank in Artemisia species: how it responds to sand mobility. Trees-struct Funct 24: 435-441.
- [23] Matthews, T. J., Cottee-Jones, H. E. W., Whittaker, R. J. (2014): Habitat fragmentation and the species–area relationship: a focus on total species richness obscures the impact of habitat loss on habitat specialists. Divers. Distrib. 20: 1136-1146.
- [24] Murphy, G. E. P., Romanuk, T. N. (2014): A meta-analysis of declines in local species richness from human disturbances. Ecol. Evol. 4: 91-103.
- [25] Phillips, H. R. P., Halley, J. M., Urbina-Cardona, J. N., Purvis, A. (2018): The effect of fragment area on site-level biodiversity. – Ecography. 41: 1220-1231.
- [26] Santos, B. A., Peres, C. A., Oliveira, M. A. (2008): Drastic erosion in functional attributes of tree assemblages in Atlantic forest fragments of northeastern Brazil. – Biol Conserv 141: 249-260.
- [27] Swihart, R. K., Gehring, T. M., Kolozsvary, M. B. (2003): Responses of 'resistant' vertebrates to habitat loss and fragmentation: The importance of niche breadth and range boundaries. Divers Distrib 9: 1-18.
- [28] Winfree, R., Fox, J., Reilly, J. R., Carveau, D. P. (2015): Abundance of common species, not species richness, drives delivery of a real-world ecosystem service. – Ecol. Lett. 18: 626-635.
- [29] Wu, J., Liu, Z. M., Qian, J. Q. (2013): Non-linear effect of habitat fragmentation on plant diversity: Evidence from a sand dune field in a desertified grassland in northeastern China. – Ecol Eng 54: 90-96.
- [30] Wu, J., Qian, J. Q., Hou, X. Z., Carlos, A. B., Liu, Z. M., Xing, B. Z. (2016): Spatial variation of plant species richness in a sand dune field of northeastern Inner Mongolia, China. J Arid Land 8: 434-442.
- [31] Zobel, M., Moora, M., Herben, T. (2010): Clonal mobility and its implication for spatiotemporal patterns of plant communities: what do we need to know next? – Oikos 119: 802-806.