

EFFECT OF VARIOUS LEVELS OF BIOCHAR INCORPORATION ON SOYBEAN YIELDS, PHYSICAL QUALITY AND NITRIFYING BACTERIA OF BLACK SOIL IN NORTHEAST CHINA

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Abstract. Maize-staw biochar (BC) was applied in a black soil area in Northeast China for investigating its effects on crop yield, soil physical properties and nitrifying bacteria. In this study, BC forms BC1, BC2 and BC3 were incorporated at levels of 15.75, 31.50 and 47.25 t ha⁻¹, respectively, from 2013 to 2014. The control soil was treated with chemical fertilizer (NPK). The results showed that yields of soybean were significantly improved overall due to BC applications, compared with the control. Meanwhile, average soil compaction values of 0 – 20 cm deep soil were decreased from 17.9% to 47.4%. Biochar supplementation decreased the soil wilting coefficients, but increased saturated water content. The available soil water content increased from 0.211 to 0.329 cm³ cm⁻³ (R² = 0.97), whereas the bulk density decreased from 1.42 to 1.40 g cm⁻³ (R² = 0.97) in response to BC supplementation. Unexpectedly, soybean yields began to decline at 31.50 and 47.25 t ha⁻¹ of BC application. as well as soil aggregate stability, available water content, and capillary pores, mainly. Moreover, the BC supplementation significantly affected soil ammonia-oxidizing archaea (AOA) abundance. These results demonstrate that BC has positive potential for enhancing crop and soil properties performances in black soil at the level of 15.75 t ha⁻¹ under continuous soybean system.

Keywords: *maize-derived biochar, application levels, physical properties, field experiment, black soil*

Introduction

Given the growth rates since 1960, the global population is expected to reach 9.7 billion by 2050 (Gonzalo et al., 2016). The consequences of human population growth have great impacts on the environment, and also on social and economic development. China accounts for approximately 22% of the world's population which is expected to increase to 14.53 billion by 2030 (Yin and Guo, 2015). Consequently, the additional food required to feed future generations would challenge the production capacity of arable land in the years to come (Golley and Zheng, 2015). The agricultural sustenance of a growing population in the black soil area of Northeast China is threatened by excessive exploitation of soil resources, leading inevitably to soil degradation. In the long run, the current production capacity of the soil may not sustain the long-term needs of a growing population under the current agricultural management level. Thus, control of soil erosion and prevention of organic matter loss have been proposed as critical strategies for sustained agriculture in this area (Luo et al., 2017; Lenka et al., 2017). In Northeast China, manure and corn straw have been used to improve soil fertility and crop yields (Hui et al., 2017; Tong et al., 2017; Liang et al., 2016). Moreover, carbon sequestration and bacterial soil contents have been improved using corn straw and manure (Yang et al., 2017; Ding et al., 2016). However, in contrast to these substrates, biochar (BC) is highly recalcitrant and unavailable to soil microorganisms (Sizmur et al., 2016).

Biochar, being a carbonaceous compound, has been used as a soil additive to counter the degradation of arable land (Luo et al., 2017). Several studies have been carried out on the effects of BC supplementation on soil aggregate formation and the physical properties of a variety of soils (Obia et al., 2016; Pratiwi and Shinogi, 2016; Lim et al., 2016; Guo, 2016). Indeed, Burrell et al. (2016) have reported that BC from straw improved the stability of soil aggregates in Planosol and Cambisol soils, and also improved the water availability to plants in Planosol soil by 38%. A soil study in the Huang-Huai-Hai Plain, China revealed that treatment with BC significantly decreased the bulk density, but increased its total porosity (Du et al., 2016). Moreover, studies have been carried out on the effect of BC supplementation level on the physical and hydraulic properties of soil. In these studies, Igalavithana et al. (2017) applied corn residue BC at levels of 2.5, 5.0, 7.5 and 10% ($w \cdot w^{-1}$) in sandy loamy soil, and after equilibration for 30 days, there were highly significant exponential reductions in K_{sat} as a result of incorporation of BC500 at levels of up to 7.5%, with K_{sat} approaching an asymptote at 10% BC500. In another study, it was shown that incubation of soil for 36 days with BC at levels of 1, 2, 5 and 10% (by wt.) reduced the water losses through evaporation and stabilized the activities of some extracellular enzymes (Elzobair et al., 2016). In field experiments, Ippolito et al. (2016) reported that graded applications of BC (1, 2, and 10% by wt.) improved the water content of calcareous soil. Supplementation of podzolic soil with BC at levels of 10, 20 and 30 $t \cdot ha^{-1}$) for 3 years produced significant and positive effects on the yield of winter rye (Kraska et al., 2016). However, most recent studies focused on low-fertility soils or under-leached, acidic saline soils. Thus, not much is known about the effect of levels of application of BC on crop yield and physical conditions of soil in degraded soils containing higher organic matter. The objective of the present study was to investigate the influence of different levels of maize straw-derived BC on continuous soybean production and soil properties such as physical environment and soil nitrifying bacteria. It was hypothesized that higher BC levels should exert positive and negative effects on soybean yield, as well as

the physical properties of black soil. Three levels of BC supplementation (15.75, 31.5 and 47.25 t ha⁻¹), each mixed with a chemical fertilizer (NPK) separately in a selected field of black soil (Mollisols), were used. Crop yield and physical and hydraulic properties of soil such as compaction, aggregation stability, bulk density, hydraulic parameters and soil surface area were analyzed. The nitrifying bacteria in the BC-supplemented soils were also evaluated.

Materials and methods

Site, soil and biochar

A long-term BC field experiment was initiated in Minzhu town, Daowai district in Haerbin city, China (E126°51'05", N45°50'3", Fig. 1). Climatological data, soil properties of surface layer and biochar properties can be seen in *Tables 1, 2, and 3* respectively. Biochar applied in this study was derived from maize straw and manufactured by Liaoning Biochar Engineering Technology Center.

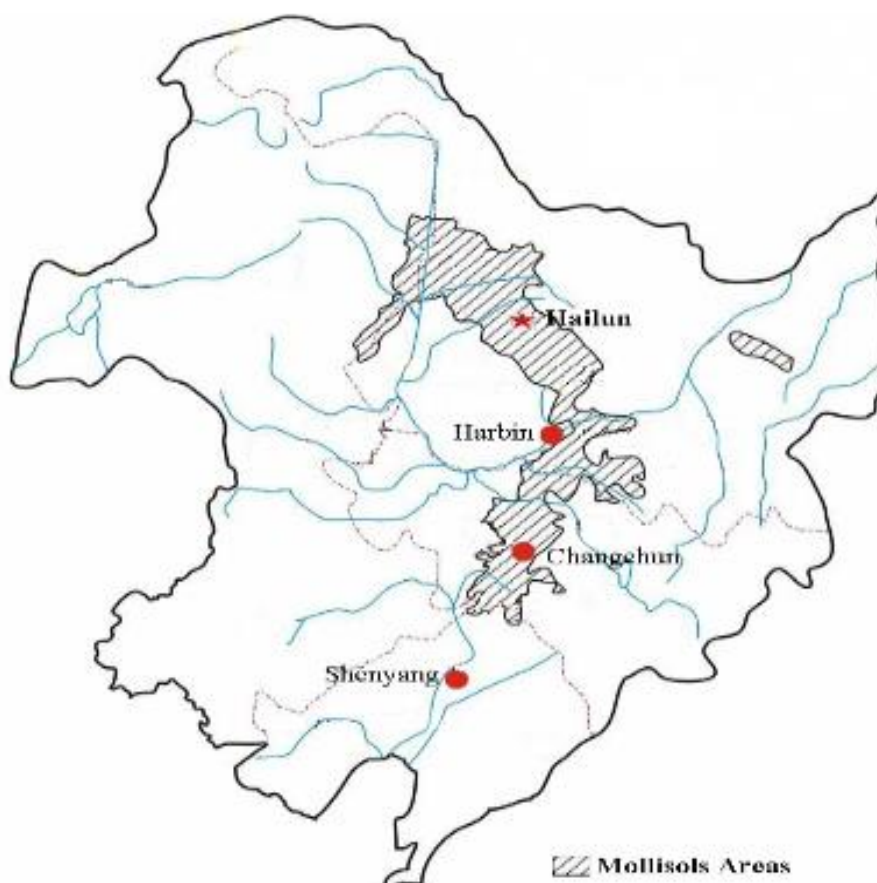


Figure 1. Map of study area

Table 1. Climatological data of experiment station

Average annual rainfall (mm)	Mean annual wind speed (m.s ⁻¹)	Maximum wind speed (m.s ⁻¹)	Sea level (m)	Ground water table depth (m)
514	4.1	18.9	138	80

Table 2. Basic chemical and physical properties in soil top layer in experiment plots

Depth cm	Mechanical composition (% America)			Texture	Available N	Available P	Available K	SOM*	pH	BD**
	Sand	Silt	Clay		mg.kg ⁻¹	mg.kg ⁻¹	mg.kg ⁻¹	g.kg ⁻¹		g.cm ⁻³
0-30	21.8	56.3	21.9	Silty clay loam	163.3	20.61	187.92	29.87	6.74	1.31

*SOM: soil organic matter; **BD: bulk density

Table 3. Biochar components

Particle components %												
SOC* g.kg ⁻¹	O g.kg ⁻¹	N g.kg ⁻¹	P g.kg ⁻¹	K g.kg ⁻¹	Si g.kg ⁻¹	Mg g.kg ⁻¹	Ca g.kg ⁻¹	pH	< 0.1 mm	0.1-2 mm	> 2 mm	
598	166	7.85	1.327	17.0	60	2	3	8.69	15.0	60.2	24.8	

*SOC: soil organic carbon

Treatments

A field experiment with soybean was carried out in 2013 and 2014. The treatments consisted of 3 levels of BC supplementation viz: NPK only, NPK + BC at level of 15.75 t.ha⁻¹ (BC1), NPK + BC at level of 31.50 t.ha⁻¹ (BC2), and NPK + BC at level of 47.25 t.ha⁻¹ (BC3). Each treatment was replicated thrice. In order to guarantee soil and plant sampling, every experiment plot was set up by covering an area of 39 m² (6 ridges×0.65 m width× 10 m ridge length). The sowing density of soybean was 300,000 plants per hectare. Soybean received sub-surface fertilizer application of 47 kg N ha⁻¹, 78 kg P₂O₅ ha⁻¹ and 68 kg K₂O ha⁻¹, with which insecticide were applied of 0.13 to 0.2 kg chlorpyrifos granules ha⁻¹ for controlling grubs. Spring soybean cultivar Heinong 58 was planted in May 2013 and harvested in October 2013, while the one planted on May 10, 2014 was harvested on October 5 in 2014. Prior to planting, the chemical fertilizer (NPK) for soybean was applied. In the 2013 period, the BC was spread over the furrow by the ridges. It was thoroughly mixed with soil using a ploughing machine (a tractor with a rototiller, Fig. 2), and then plowed to a depth of over 20 cm and then mainly ranked for a leveling. Herbicide applied over the ridges just after soybean sowing (Table 4).

Table 4. Basic physical and chemical properties of soil after harvest in 2014

Treatment	N (%)	H (%)	O (%)	C (%)	BD* g.cm ⁻³	CEC** cmol.kg ⁻¹	SSA*** m ² .g ⁻¹	C/N	O/C	H/C
NPK	0.094 ± 0.01d	0.666 ± 0.01a	2.290 ± 0.09b	0.915 ± 0.01c	1.40 ± 0.05a	22.77 ± 0.23c	29.35 ± 0.43d	9.76	2.50	0.73
BC1****	0.179 ± 0.02c	0.624 ± 0.03a	2.344 ± 0.13a	1.800 ± 0.02a	1.28 ± 0.03b	23.08 ± 0.41b	32.74 ± 0.61a	10.05	1.30	0.35
BC2	0.186 ± 0.01b	0.660 ± 0.02a	2.211 ± 0.21b	1.150 ± 0.03b	1.18 ± 0.02d	24.72 ± 0.53a	31.62 ± 0.21b	6.17	1.92	0.57
BC3	0.203 ± 0.02a	0.647 ± 0.01a	1.944 ± 0.11c	0.885 ± 0.02d	1.22 ± 0.02c	24.10 ± 0.20a	30.19 ± 0.20c	4.37	2.20	0.73

BD*: bulk density; CEC**: cation exchange capacity; SSA***: specific surface area; BC****: biochar



Figure 2. Ploughing machine working in the biochar-additive field plots

Measurements

Particle size distribution of BC

The particle size distribution of BC was determined by dry-sieving the samples using a sieve shaker (Endecott Test Sieve Shaker, Watson Victor Ltd.). Seven different fractions were obtained using 2.00, 1.00, 0.50, 0.25 and 0.106 mm sieves. Three consecutive shakings were conducted, and it was observed that the weights of different fractions remained unchanged. The first shaking was continued for 3 min, and the other two shakings were done for only 2 min. Measurements of N₂ gas adsorption for the determination of BET (Brunauer-Emmett-Teller) surface area of BC were carried out using a Micromeritics ASAP 2020 volumetric adsorption system.

Soil compaction

In the field, the soil compaction values were measured with in 2014 after the harvest of soybean with a soil compaction meter (SC-900, *Fig. 3*). The data for the top layer (0-20 cm) were automatically recorded once every 2.5 cm as the soil compaction meter descended from the surface.



Figure 3. Soil compaction meter (SC-900)

Soil aggregates and their stability parameters

The size distribution of wet-sieving aggregates was determined with a rototap machine containing a nest of eight 100-mm diameter sieves with screen openings of 10, 7, 5, 2, 1, 0.5 and 0.25 mm. The weight of soil retained on each sieve was measured after dry-sieving 0.5 kg of soil for 2 min. Moreover, 50 g of soil aggregates retained on 100 mm sieve was used for the analysis of wet-sieving which was carried out mechanically using Yoder's apparatus.

The nest of sieves used had mesh sizes of 2, 1, 0.5, 0.25 and 0.106 mm. Wet sieving was continued for 20 min with 40 oscillations per min. The aggregates retained on each sieve were transferred to a set of pre-weighed beakers, oven-dried at 105 °C for 24 h,

and weighed. The mean weight diameter (MWD) and geometric mean diameter (GMD) were calculated as indices of aggregation as shown in *Equations 1* and *2*:

$$MWD = \frac{\sum_{i=1}^n (\bar{d}_i W_i)}{\sum_{i=1}^n W_i} \quad (\text{Eq.1})$$

$$GWD = \text{EXP} \left[\frac{\sum_{i=1}^n W_i \ln(\bar{d}_i)}{\sum_{i=1}^n W_i} \right] \quad (\text{Eq.2})$$

where \bar{d}_i is the mean diameter of the class (mm), and W_i is the proportion of aggregate retained on the sieves. The fractal dimension was determined in line with method of Yang et al. (1993), as shown in *Equation 3*:

$$\frac{M(r < \bar{X}_i)}{M_T} = \left(\frac{\bar{X}_i}{X_{\max}} \right)^{3-D} \quad (\text{Eq.3})$$

where $M(r < \bar{X}_i)$ is the cumulative mass of objects or fragments of the i^{th} size, $r < \bar{X}_i$; M_T is the total mass of particles; \bar{X}_i is the mean particle diameter (mm) of the i^{th} size class; and X_{\max} is the mean diameter of the largest particle.

Hydraulic parameters

The water retention curve θ (h) of soil was determined using the pressure plate method. Saturated hydraulic conductivity (Ks) was determined using a fixed-head permeameter instrument, and the parameters of θ (h) and unsaturated hydraulic conductivity K (h) were obtained using the van Genuchten–Mulem model (*Eq. 4*).

$$\theta(h) = \theta_r + \frac{\theta_s - \theta_r}{[1 + |\alpha h|^n]^m} \quad (\text{Eq.4})$$

This equation was fitted on the moisture retention data of soil, and the parameters i.e. saturated water content (θ_s), residual water content (θ_r), inverse of suction at the inflexion point of the moisture curve (α), and shape parameters (m and n) were estimated through a non-linear least-squares optimization using RETC (RETention Curve) software.

Pore size distribution

The soil pore size distribution (PSD) was determined using mercury intrusion porosimetry (MIP) (Autopore IV 9500, Micromeritics Inc. USA). In the MIP method, the mercury pressure was increased stepwise, and the intruded volume of mercury was monitored for each pressure in the range of 0.0036–310 MPa. The MIP test indicates the volume of cumulative mercury intruded as a function of equivalent pore radius (EPR). The results obtained were plotted in two graphical forms: cumulative pore volume versus logarithmic EPR, and differential PSD versus logarithmic differentiation (dV/dlog r). The values of pore radii on the cumulative curve and differential curve were 0.003 μm and 360 μm , respectively. This wide range allowed for the detection of diverse soil pore classes along the PSD curve. The pores were classified according to their equivalent pore diameter (EPD) into five classes of sizes: macropores (>75 μm),

mesopores (30–75 μm), micropores (5–30 μm), ultramicropores (0.1–5 μm), and cryptopores (0.1–0.007 μm) according to the method of Camerson and Buchan (2006).

Soil ammonia bacteria [ammonia-oxidizing archaea (AOA) and ammonia-oxidizing bacteria (AOB)]

Total DNA was extracted from frozen soil samples (0.5 g wet weight) using a FastDNA® SPIN Kit For Soil (MP Biomedicals, LLC) in line with the manufacturer's instructions. The concentration and quality of the extracts were determined using UV spectrophotometer (Implen, München, Germany), and the extracts were kept at $-20\text{ }^{\circ}\text{C}$ prior to further molecular analysis. Real-time PCR assays (total reaction volume = 20 μL) were carried out using SYBR®Premix Ex Taq™ Perfect Real Time (Takara, Dalian, China), and multicolor real-time PCR detection system (Bio-Rad Laboratories Inc., Hercules, CA, USA). The primer sets (AOA: Arch-amoAF/Arch-amoAR; AOB: amoA-1F/amoA-2R) and thermal profiles used to amplify each target gene with real-time PCR are listed in *Table 1*. The abundance (copy number) of bacterial AOA and AOB genes were calculated using a regression equation to convert the cycle threshold (Ct) values to the known number of copies in the standard curves.

Statistical analysis

All data were analyzed using Microsoft Excel (2003). The significance of differences among different treatments and sampling dates were tested with ANOVA using SPSS software package (SPSS Inc., 2003). Differences between values were considered statistically significant at $p < 0.05$. The coefficient of determination (R^2) of non-linear regression was used to determine the best fit of the water retention model of soil.

Results

Site, soil and BC

The effect of BC application on soybean yields (2013 and 2014) is shown in *Figure 4*. Soybean yields were significantly influenced by BC over two years, when compared with the control (NPK). In the first year, BC1, BC2 and BC3 increased soybean yields by 18.5, 12.5 and 9.7%, respectively, when compared with NPK. Similarly, in the next year, increases in the soybean yield ranged from 15.4% (for BC1) to 4.6% (for BC3). The predominant yield for each year was produced by BC1.

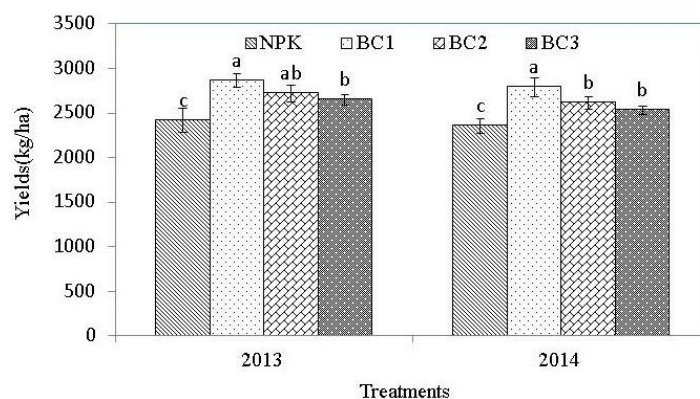


Figure 4. Effect of BC application on soybean yields from 2013 to 2014

Effect of field application of BC on the distribution of soil compaction in the top layer (0-20 cm)

Soil compaction reflects the conditions of pores in the soil and the strength of the junction force in the soil particles. After two years of field application of BC, there was an ‘S’ trend in the distribution of soil compaction values from the top of the soil (0–20 cm layer) for all treatments, and the compactability increased with increase in soil depth (Fig. 5A).

Compared with NPK treatment, BC1, BC2 and BC3 decreased average values of soil compactability distribution within 0-20 cm of soil by 47.4, 38.4 and 17.9%, respectively. The order of average soil compaction values was NPK > BC3 > BC2 > BC1. Except for NPK treatment, the strength of soil compaction was positively related to the supplementation level of BC i.e. BC3 > BC2 > BC1.

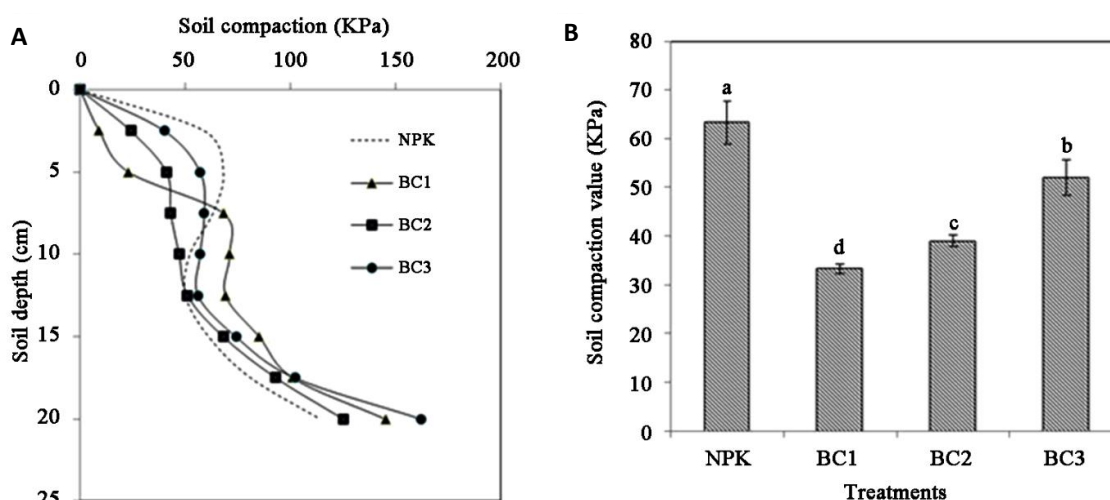


Figure 5. Effect of BC on the distribution of soil compaction in the top soil layer (0-20 cm, (A)); average soil compaction values at 0-20 cm (B)

Effect of field application levels of BC on the water-stable aggregates

The details are presented in Table 5.

Table 5. Effect of level of supplementation of BC on the composition and stability indices of water-stable aggregates for two years

Treatments	Aggregate size composition (%)				> 0.25 mm Aggregates (%)	MWD* (mm)	GWD** (mm)
	> 2 mm	2-0.25 mm	0.25-0.106 mm	< 0.106 mm			
NPK	4.54 ± 0.21c	57.94 ± 1.17c	21.92 ± 1.35b	15.60 ± 0.61b	62.48 ± 2.61c	0.26 ± 0.05c	0.27 ± 0.01a
BC1	7.60 ± 0.41a	62.68 ± 1.05a	19.22 ± 1.89c	10.50 ± 0.25d	70.28 ± 1.43a	0.33 ± 0.02a	0.27 ± 0.01a
BC2	6.40 ± 0.11b	60.76 ± 2.06b	19.72 ± 0.94c	13.12 ± 1.54c	67.16 ± 1.51b	0.30 ± 0.01b	0.28 ± 0.02a
BC3	6.19 ± 0.18d	49.01 ± 1.03d	26.02 ± 0.61a	18.78 ± 0.92a	55.20 ± 0.85d	0.27 ± 0.04c	0.28 ± 0.01a

*MWD = mean weight diameter; **GWD = geometric mean diameter

Aggregate and its stability

The results from wet-sieving (*Table 5*) showed that the addition of BC significantly affected the amounts of macroaggregates (> 0.25 mm), which were increased by 12.5, 7.5 and -11.7% for BC1, BC2 and BC3, respectively, when compared to control (NPK). The changes in the aggregate sizes of 2-0.25 mm showed a positive relationship with MWD values with a correlation coefficient of 0.64 ($P < 0.05$). The mean weight diameter (MWD), geometric mean diameter (GMD) can be used for amended-soil aggregates in the evaluation of the aggregation stability of soil. Higher values of these indices indicate the predominance of the more stable aggregates over the smaller, less stable fractions (Nath and Rattan, 2017; Rabot et al., 2018). From *Table 5*, after two years of the field application of BC in 2014, the MWDs of the soil aggregate of the three treatments (BC1, BC2 and BC3) were enhanced by 26.6, 15.4 and 3.8%, respectively over the control. No significant differences in GMD were found among the four treatments.

Effect of BC on the hydraulic characteristics of soil

Hydraulic parameters were significantly influenced by the level of BC incorporation, except in a few items in this study such as θ_s (for BC1, BC2 and BC3) and total porosity (for BC2 and BC3) compared with the control (*Table 6*). Compared to NPK, the incorporation of maize BC at levels of 15.75, 31.5 and 47.25 t.ha⁻¹ increased field capacity to 0.379, 0.381 and 0.326 cm.cm⁻¹, respectively, from initial value of 0.291 cm.cm⁻¹. As shown in *Table 6*, wilting coefficient (θ_r) values were significantly reduced, while saturated water content (θ_s) values were not significantly increased at all levels of BC treatments ($p < 0.05$). Smaller particle fraction in BC increased water holding space in black soil more significantly than larger particles (*Tables 2 and 5*). The results showed clearly that incorporation of BC at the level of 15.75 t.ha⁻¹ increased the soil k_s , θ_s and available water content by 50.14, 17.94 and 55.92%, respectively, when compared to NPK control ($p < 0.05$). The available water content refers to the difference between field capacity and wilting point; it is the quantity of soil water available for the usage of plants. The highest available water content was 0.332 cm.cm⁻¹ for the soil containing BC at the level of 15.75 t.ha⁻¹, while BC at levels of 31.5 and 47.25 t.ha⁻¹ decreased available water content to 0.316 and 0.266 cm.cm⁻¹, respectively. The regression equation for the available water capacity of soil as a function of level of incorporation of BC is shown in *Equation 5*:

$$\text{Available water content} = -0.0002(\text{t ha}^{-1}\text{biochar})^2 + 0.0126(\text{t ha}^{-1}\text{biochar}) + 0.2684 \text{ (Eq.5)}$$

It is clear from the quadratic equation that the addition of BC at the level of 15.75 t.ha⁻¹ increased available water content in the soil, but decreased it at levels of 30.5 and 47.25 t.ha⁻¹ ($R^2 = 0.9355$). Field capacity was enhanced by BC incorporation, reaching the highest value of 0.382 cm.cm⁻¹, but wilting point was significantly lower with BC3 than BC2. Thus, the difference (available water content) determined by the two coefficients for BC level of 15.75 t.ha⁻¹ was lower (0.329) than that for BC level of 31.5 t.ha⁻¹ (0.328).

Figure 6 shows the regression models developed from the results on *Table 6*, which indicates the positive effect of BC on bulk density and available water content. The lower bulk density of the biochar-supplemented soil resulted directly from the

contribution of pores of BC, since the bulk density of BC-excluded treatment was significantly lower than that of the unamended NPK control. This may be due to changes in total porosity. The results of hydraulic conductivity revealed an initial increase at BC level of 15.75 t.ha⁻¹, which was followed by decreases at BC supplementation levels of 31.5 and 45.25 t.ha⁻¹. This may be due to the hydrophobic nature of organic matter in BC.

Table 6. Effect of BC on hydraulic parameters of soil samples

Treatment (0-30 cm)	Θ_r^*	Θ_s^{**}	α^{***}	n^{***}	k_s^{****} (mm.min ⁻¹)	Field capacity (cm.cm ⁻¹)	Available water content (cm.cm ⁻¹)	Total porosity %
NPK	0.080 ± 0.005a	0.340 ± 0.03a	0.078 ± 0.01a	1.96 ± 0.17b	10.71 ± 0.56d	0.291 ± 0.02d	0.211 ± 0.02d	47.16 ± 2.61d
BC1	0.050 ± 0.003b	0.401 ± 0.01b	0.017 ± 0.01d	1.71 ± 0.09c	16.08 ± 0.42a	0.379 ± 0.03b	0.329 ± 0.03c	55.47 ± 1.51a
BC2	0.056 ± 0.002c	0.397 ± 0.02b	0.035 ± 0.01b	2.28 ± 0.15a	15.48 ± 0.23b	0.381 ± 0.01a	0.325 ± 0.02a	53.96 ± 0.49b
BC3	0.060 ± 0.005d	0.389 ± 0.02b	0.023 ± 0.02c	1.05 ± 0.3d	11.90 ± 1.05c	0.326 ± 0.03c	0.266 ± 0.02b	51.70 ± 0.84b

* θ_r = soil residual water content/wilting coefficient (cm³.cm⁻³); ** θ_s = soil saturated water content (cm³.cm⁻³); *** α and n = parameters related to the shape of the soil water characteristics curve; **** k_s = hydraulic conductivity

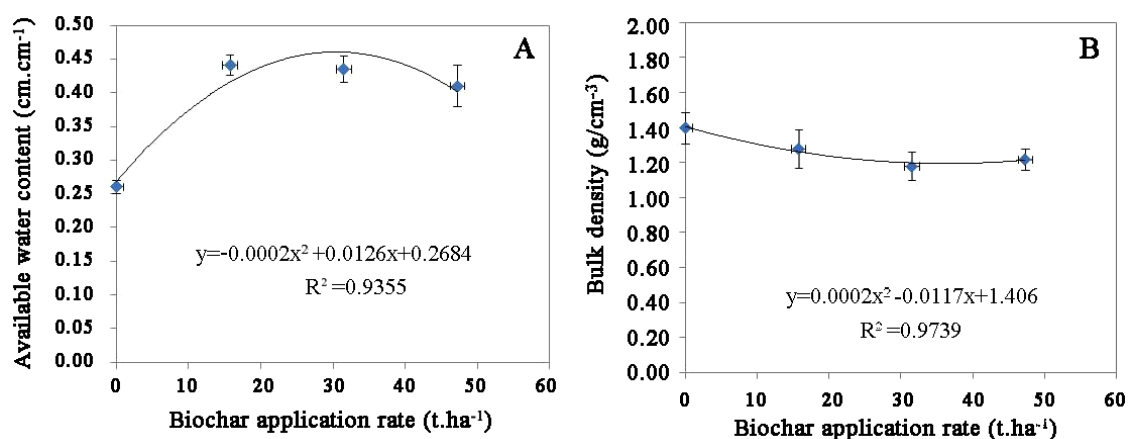


Figure 6. Regression models for the physical properties and the level of incorporation of BC. A: model for available water content and level of BC supplementation; B: model for bulk density and level of incorporation of BC

Effect of BC on the pore distribution of soil

The results of pore analysis porosity and pore size distribution of BC-amended soils are shown in Table 7. Total pore volume and porosity of soils were higher ($p < 0.05$) in BC-treated soils than in NPK control soil. The degree to which the total porosity values increased with the BC incorporation level varied among the treated soils. The relative increases in total porosity due to BC1, BC2 and BC3 were 17.6, 14.4, and 9.6%, respectively, relative to NPK control. The BC treatment caused significant changes in

the pore in the ranges of > 75, 30-75, 5-30, and 0.1-5 μm , which indicates that it indeed altered the macro-, meso-, and micropores. However, BC did not significantly affect the cryptopores (< 0.1 μm). In SB-amended soil, approximately 32% of the total porosity was in > 75 μm pores, followed by 0.1-5 μm class (approximately 25%), 5-30 μm class (approximately 14%), and 0.01-0.1 μm class (approximately 8%). The changes in pore size distribution brought about by BC treatments could possibly reflect the particle size distribution of BC. Compared with BC1, BC at the higher level of 31.5 $\text{t}\cdot\text{ha}^{-1}$ increased the total volume, but decreased the macropores and mesopores by 53.54 and 16.50%, respectively, indicating that BC2 produced more unavailable pores leading to less available water in the soil. Compared with BC2, the 78.76% increase in 5-30 μm micropores, and 74.42% decrease in 30-75 μm mesopores are in agreement with the hypothesis that direct changes in capillary pores are caused by the occupation of finer BC particles, leading to decreased available water content at BC levels of 31.5 and 47.5 $\text{t}\cdot\text{ha}^{-1}$.

Table 7. Total porosity and pore-size distribution of BC-amended soils

Treatment	Total volume (cm^3g^{-1})	Pore size distribution (cm^3g^{-1})						
		Total porosity (%)	> 75 μm	30-75 μm	5-30 μm	0.1-5 μm	0.1-0.01 μm	< 0.01 μm
NPK	0.2559 \pm 0.01d	47.16 \pm 1.52d	0.0975 \pm 0.03b	0.0234 \pm 0.01c	0.0365 \pm 0.0028c	0.0600 \pm 0.0041c	0.0261 \pm 0.0017c	0.0124 \pm 0.0005b
BC1	0.4115 \pm 0.02b	55.47 \pm 0.97a	0.1892 \pm 0.05a	0.0988 \pm 0.03a	0.0224 \pm 0.0019d	0.0598 \pm 0.0057c	0.0288 \pm 0.0011b	0.0125 \pm 0.0002b
BC2	0.4201 \pm 0.01a	53.96 \pm 1.12b	0.0879 \pm 0.04d	0.0825 \pm 0.01b	0.053 \pm 0.0049b	0.1389 \pm 0.0093a	0.0455 \pm 0.0035a	0.0123 \pm 0.0004b
BC3	0.3197 \pm 0.01c	51.70 \pm 2.51c	0.0910 \pm 0.01c	0.0211 \pm 0.02d	0.0771 \pm 0.0065a	0.0902 \pm 0.065b	0.0271 \pm 0.0012bc	0.0132 \pm 0.0003a

Effect of BC incorporation level on the soil nitrifying oxidizers

Changes in soil nitrifying bacteria [soil ammonia-oxidizing archaea (AOA) and soil ammonia-oxidizing bacteria (AOB)] of BC-treated soils are presented in *Table 8*. Values of AOA and AOB ranged from 0.91 to 1.16 $\times 10^8$ $\text{cfu}\cdot\text{g}^{-1}$, and from 0.65 to 0.76 $\times 10^7$ $\text{cfu}\cdot\text{g}^{-1}$, respectively.

Table 8. The abundance of soil nitrifying bacteria of biochar-amended soils during harvest in 2014

Treatment	AOA (10^8)	AOB (10^7)	pH
	$\text{Cfu}\cdot\text{g}^{-1}$	$\text{Cfu}\cdot\text{g}^{-1}$	
NPK	1.09 \pm 0.03b	0.65 \pm 0.03b	6.80 \pm 0.23a
BC1	1.16 \pm 0.02a	0.81 \pm 0.02a	7.13 \pm 0.51b
BC2	0.97 \pm 0.04c	0.73 \pm 0.02b	7.29 \pm 0.34c
BC3	0.91 \pm 0.05cd	0.76 \pm 0.04ab	7.65 \pm 0.13b

Similar to pH, AOA abundance was significantly enhanced by the addition of BC, when compared to NPK, but AOB was not significantly increased. The ratio of AOB to

AOA increases with increasing pH gradient in the rhizosphere (Yu et al., 2019). Biochar at the level in BC1 enhanced AOA and AOB copies, but decreased them in BC2 and BC3.

Discussion

The application of BC enhances crop yields as well as a range of physical properties of soil (Laird et al., 2017; Agegnehu et al., 2016; Burrell et al., 2016; Du et al., 2016). In addition, BC influences soil nitrifying bacteria (Pan et al., 2017; Radwan et al., 2018; Azerang et al., 2018). The application of BC significantly increased the yield of soybean due to soil structure and improvement in moisture content as a result of changes in soil nitrifying bacteria. This can be attributed to the enhanced specific surface area and higher porosity as seen in the present study, which have been shown to enhance microbial activity and increase SOM (Chen et al., 2017; Ren et al., 2018; Sriprechasak et al., 2018; Sangdee et al., 2017). When BC is applied under the influence of tillage and cultivation operations, the distribution of soil compaction in the top layer of soil could be due to a dilution effect for the lighter fraction at the top layer (0-20 cm) (Soane, 1990). It has been reported that microaggregate formation and aggregate stability are significantly improved by the addition of BC (Kaiser et al., 2017; Hartley et al., 2016). The macroaggregates of black soil were significantly increased by incorporation of maize straw-derived BC at the levels of 15.75 and 31.5 t.ha⁻¹, and the stability of BC1 (15.75 t.ha⁻¹) was enhanced in the additive treatments. On the other hand, BC at levels of 31.5 and 47.25 t.ha⁻¹ produced negative effects on aggregate formation, leading to decreases in percentage of macroaggregates and lower MWD values, relative to BC1. This may be partially due to the fine particles with excessive biochar occupying the exist capillary under the external forces of mechanical cultivation and rainfall (Busscher et al., 2010) and then the restrict of microbial activity and nitrogen mineralization (Dempster et al., 2012) leading to less soil cement existed in unit soil mass. Another reason could be the effect of excessive BC which effectively decreased the risk of surface sealing in the BC-amended soils, relative to BC1 (Sun and Lu, 2014). Moreover, the C:N ratio of BC1 was the highest (10.05, *Table 4*), whereas the O:C and H:C ratios were the lowest (1.32 and 0.35, respectively, *Table 4*) among the four treatments (*Table 4*). Generally, a critical C:N ratio of < 25 is important because it indicates the probability of net mineralization of nitrogen (Chapin et al., 2002). The stability of BC-amended soil with a dominant level of BC could also be reflected in O:C and H:C ratios, the correlation coefficient of which was positive 0.97. The lower values of the two ratios resulted in a more stable BC-amended soil (Spokas, 2010; Budai et al., 2013), which could be a major factor that influences the formation of soil aggregates.

The effect of application of BC on water retention capacity is linked to the improvement of soil aggregation or structure (Ibrahim et al., 2017). The available water content in the soil was increased by BC addition at the level of 15.75 t.ha⁻¹, but was decreased at BC levels of 31.5 and 47.25 t.ha⁻¹ mainly due to the changes in the capillary pores of the BC-amended soil. The negative effect of these changes could be due to the partial replacement of the clay particles by the finer particles in the BC, followed by a decrease in wilting point. Studies by Nadeem et al. (2017) showed that BC enhanced field capacity, which is consistent with the outcomes of this study. In a series of well constrained laboratory experiments, Liu et al. (2017a) showed that BC particle sizes affected soil water content by changing the pore spaces between particles

(interpores), and by adding pores that are part of the biochar (intrapores). Moreover, it has been reported that BC treatment either decreases (clogs) or increases pore spaces in the mixture, based on the quantity of fine BC fraction, which in turn could decrease or increase the hydraulic conductivity of the mixture (Trifunovic et al., 2018).

The addition of excess amount of BC ($15.75 \text{ t}\cdot\text{ha}^{-1}$) probably led to filling up of some of the capillary pores existing in the BC-amended soil by the additional dust ($< 0.1 \text{ mm}$) that accompanied BC. Thus, the water holding capacity of the soil was reduced at higher BC levels. It has been reported that fine particles of excessive BC fill the soil pores, thereby reducing the permeability while increasing the water retention of BC-amended soil (Uzoma et al., 2011; Busscher et al., 2010). This provides the physical environment in response to changes in soil nitrifying bacteria. Similar to the available water content of soil, bulk density is a parameter which is useful for measuring the relative mass of a solid relative to the bulk volume that the solid occupies, including void spaces. For the non-amended soil, the bulk density was $1.40 \text{ cm}\cdot\text{cm}^{-1}$, and was decreased to 1.28, 1.18, and $1.22 \text{ g}\cdot\text{cm}^{-3}$ by BC1, BC2 and BC3, respectively. Bulk density also showed a quadratic regression relationship with increasing level of BC ($R^2 = 0.97$). Recent studies on the negative effect of high BC application on soil-plant system revealed that addition of BC at levels of 10, 20 and $30 \text{ t}\cdot\text{ha}^{-1}$ produced positive influence on grain yield of winter rye, with $20 \text{ t}\cdot\text{ha}^{-1}$ resulting in the highest grain yields (Kraska et al., 2016). Pot experiments revealed that *S. salsa* yield increased from 11.7 to 115% with WS application in the range of $5\text{--}10 \text{ g}\cdot\text{kg}^{-1}$, when compared with control (Sun et al., 2016). However, as BC level increased to $20 \text{ g}\cdot\text{kg}^{-1}$, the yield decreased to 102%. Usowicz et al. (2016) have reported that incorporation of BC under the fallow led to reduction of soil bulk density and particle density from $1.18\text{--}1.20 \text{ Mg}\cdot\text{m}^{-3}$ and $2.48\text{--}2.55 \text{ Mg}\cdot\text{m}^{-3}$ from $0 \text{ t}\cdot\text{ha}^{-1}$ and $10 \text{ t}\cdot\text{ha}^{-1}$ to $1.00 \text{ Mg}\cdot\text{m}^{-3}$ and $2.20 \text{ Mg}\cdot\text{m}^{-3}$ under $30 \text{ t}\cdot\text{ha}^{-1}$, respectively. In a field experiment conducted by Liu et al. (2017b), it was found that BC supplementation at levels of 10 and $20 \text{ g}\cdot\text{kg}^{-1}$ soil increased peanut nitrogen fixation by 15.52 and 14.11%, respectively, in an intercropping system where maize (*Zea mays* L.) was intercropped with either soybean (*Glycine max* L.) or peanut (*Arachis hypogaea* L.).

The results obtained in this study reveal that BC exerts both positive and negative effects on the physical environment of black soil, and provide a real growth condition for crop yield, depending on the level of application. Higher concentrations of AOA and AOB improve the conversion of other forms of nitrogen to available nitrogen fertilizer so as to enhance the fertility of soil and promote plant growth (Luo and Lin, 2013). The bacterial indices AOA and AOB are the real driving forces, with AOA being more sensitive to changes in soil rhizosphere environment, especially pH. Indeed, AOA may be regarded as a factor for direct assessment of the benefits of BC, which is in agreement with the results observed in a hydroagric Stagnic Anthrosol and an entic Halpudept (Lin et al., 2017). In this study, increases in pH occurred along with increases in AOB: AOA ratio, leading to increased nitrogen mineralization by stimulating microbial activity (Teutscherova et al., 2017). On the other hand, AOA was reduced at higher BC levels mainly due to adverse physical condition of soil, especially decreased availability of water and capillary pores, which negatively affected rhizosphere soil nutrient supply required for crop growth.

Conclusion

Maize-derived BC incorporation increased the yields of soybean, and improved the physical properties of black soil such as aggregation stability, water holding capacity and pore size distribution. It is evident that as the level of application of BC increased to 37.5 and 47.25 t ha⁻¹, soil aggregation became less stable, possibly due to the negative effect on microbial habitat, leading to decreased stability of soil aggregates and lower levels of binding agents per unit of soil volume. Similarly, available water holding capacity started to decrease at BC level of 37.5 t ha⁻¹, due mainly to the filling of the capillary pore space of BC-soil mixture with finer ashes in excess BC. Moreover, AOA and AOB were enhanced by the lowest BC level (BC1), but inhibited at higher levels (BC2 and BC3), which resulted in direct reduction of crop yield by affecting the nitrogen supply capacity of rhizosphere soil. It is necessary that these results are verified in long-time field observations. Therefore, it can be concluded from these results that the maize straw-derived BC can be applied to black soils as a strategy for optimization of sustainable agricultural production.

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