

PHYTOPLANKTON COMMUNITY STRUCTURE AND ITS RELATIONSHIP WITH ENVIRONMENTAL FACTORS IN TAIHU NATIONAL WETLAND PARK IN NORTHEAST CHINA

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Abstract. In order to understand the community structure of phytoplankton and its relationship with environmental factors in Taihu National Wetland Park in northeast China, a study was conducted in Taihu in September 2020. The results showed that there were 39 species of phytoplankton belonging to 27 genera and 6 phyla, among which Chlorophyta was the most abundant (20 species). The abundance and biomass of 5# and 6# sampling sites were significantly higher than other sampling sites, which may be caused by eutrophication. In Taihu National Wetland Park, the dominant species were *Merismopedia minima*, *Aphanizomenon flosaquae*, *Synedra acus*, *Synura*.sp and *Scenedesmus quadricauda*. Phytoplankton community structure in autumn in Taihu National Wetland Park were mainly influenced by pH, NO₃⁻-N, and water temperature. This research can provide a scientific basis for biodiversity protection and pollution control in Taihu National Wetland Park.

Keywords: *phytoplankton, biodiversity, richness, correlation, spatiality*

Introduction

According to Chalar (2009), the Earth's ecosystem is experiencing an unprecedented rate of biodiversity loss as a result of global climate change, eutrophication, and overexploitation of natural resources. Aquatic ecosystems especially lakes are vulnerable and the loss of biodiversity may have catastrophic consequences, such as algal blooms (Cardinale et al., 2012; Weyhenmeyer et al., 2013). Phytoplankton, which forms the base of food web in lakes, is particularly sensitive to variations in environmental factors (Mulvenna et al., 2012; Weyhenmeyer et al., 2013). Their community structure-like composition, abundance and biomass are affected by biotic factors, as well as abiotic factors like temperature, salinity, pH, nutrients and electrical conductivity (Oduor and Schagerl, 2007; Schagerl and Oduor, 2008; Oren, 2011). Investigations of phytoplankton communities in different ecosystem types have been carried out by different researchers in China (Ma and Yu, 2013; Sun et al., 2019; Ma et al., 2021).

Taihu National Wetland Park is located in the east of Tailai County. The wetland park receives water from direct precipitation. It also receives inflow from domestic sewage and industrial wastewater from Tailai County (Lin, 2002). Before 1990's, the wetland

park was seriously polluted because of the inflow of wastewater from Tailai paper mill. However, after the paper mill was banned from discharging its wastewater into the wetland, the water quality of Taihu National Wetland Park improved to a certain extent. In the recent years, the development of tourism and the influence of nearby residents' activities as a result of increase in human population have led to deterioration of the water quality in the wetland. As it currently stands, the water quality of the lake is in deplorable state. Although studies on planktons have been conducted in most of the aquatic systems in China (Sun et al., 2010; Yang et al., 2011; Shi et al., 2015), work of community structure of phytoplankton and their relationship to environmental variable of Taihu National Wetland Park is, however, lacking. So, this study will help fill the lacunae in scientific understanding of the community structure of this unique ecosystem. Moreover, this study will provide some reference for ecological restoration and management of the Wetland Park.

Materials and methods

Study area and sampling sites

The study was conducted in Taihu National Wetland Park which lies on E123°25'14.17"~E123°29'00.97", N46°24'12.10"~N46°21'15.10" in northeast China (Chen et al., 2018). Taihu National Wetland Park covers an area of approximately 1365 hm² and it lies in an area which is under the influence of temperate continental monsoon climate zone. This study area receives an annual average precipitation of 360 mm with an annual mean temperature of about 4.2 °C. The wetland is important for water supply, flood control, tourism and recreation, and aquaculture.

All samples were collected 3 times from 8 sampling sites of Taihu National Wetland Park in September (autumn) in 2020. The site included 1# located in the ditch close to Taihu, sites 2# and 3# were located in the sewage treatment plant, site 4# was located in the "reclaimed water" inlet of the sewage treatment plant, and sites 5-8# were located in the middle of the Taihu National Wetland Park (*Table 1, Figure 1*).

Table 1. Location of sampling sites in Taihu National Wetland Park

Sampling sites	Coordinate	Sampling sites	Coordinate
1#	46°21'56"N123°26'32"E	5#	46°23'03"N123°26'14"E
2#	46°22'18"N123°26'42"E	6#	46°23'04"N123°26'20"E
3#	46°22'25"N123°26'15"E	7#	46°23'02"N123°27'13"E
4#	46°22'20"N123°26'50"E	8#	46°23'30"N123°26'51"E

Sample collection and analysis

At each sampling site water temperature (WT), pH, ammonia nitrogen (NH₄⁺-N) and nitrate nitrogen (NO₃⁻-N) were measured using a Multi parameter water quality analyzer (YSI 6600, YSI Inc., USA) while water transparency (SD) was determined using a 20 cm Secchi disk. Whole water samples from the same site were collected and transported to the laboratory for analysis of total nitrogen (TN), total phosphorus (TP) and chemical oxygen demand (COD_{Cr}). Total nitrogen (TN), TP and COD_{Cr} were analyzed according to the standard methods for China (MEP (Ministry of Environmental Protection, 2002)).

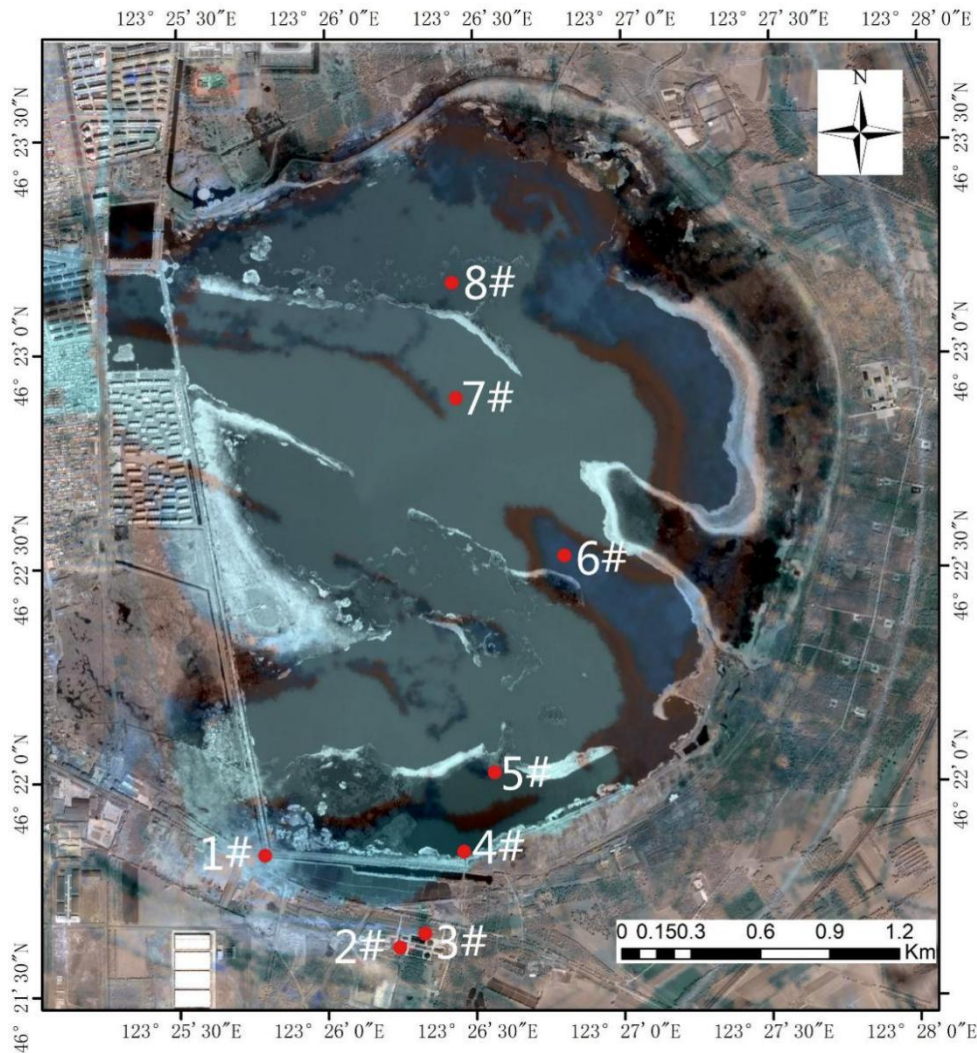


Figure 1. Map of Taihu National Wetland Park Showing the distribution of sampling sites

Replicates of phytoplankton samples (1 L at each site) were collected, put in a labeled bottle and immediately fixed with Lugol's solution. The phytoplankton samples were allowed to sediment for 48 h and then concentrated to 30 mL. Phytoplankton were identified by referring to the identification key of Hu (2006) and counted using an inverted microscope (Leica Microsystems, German) at 400× magnification. Phytoplankton diversity, richness and dominance were determined for each sampling sites using number of taxa, total number of individuals per taxa and relative abundance of each taxon over the sampling period respectively. Biomass of the phytoplankton was computed by dividing wet weight (mg) obtained from length–weight relation of the species to the volume of water (L) filtered (Sun et al., 2010). Shannon-Weaver diversity index (Shannon and Weiner 1948) was used to calculate diversity as follows:

$$H = -\sum\left[\left(\frac{n}{N}\right) * \left(\ln \frac{n}{N}\right)\right] \quad (\text{Eq.1})$$

where n = number of individuals of a taxon; N = total number of individual in a site. This index takes into account both the numbers of taxa and their relative abundance.

Taxon richness was calculated using Margalefs Index (D) (Clarke and Warwick, 1994):

$$D = (S - 1) / \ln N \quad (\text{Eq.2})$$

where S is the total number of species, N is the total number of individuals of all species, and P_i is the proportion of the number of individuals of the i-th species in the total number of individuals.

Species dominance was computed according to the formula proposed by as follows:

$$y = f_i \times P_i \quad (\text{Eq.3})$$

where, y is the dominance, P_i is the proportion of the i-th species in the total number of individuals, f_i is the frequency of the species at each sampling time. When $y > 0.02$, it is defined as the dominant species. Table 2 below shows the water quality evaluating criteria based on the diversity indices.

Table 2. Diversity Index Water Quality Evaluation Standard

Index	Pollution index		
Shannon-Weaver Diversity index(H')	0-1 Heavy	1-3 Medium	>3 Light
Margalef Richness index(D)	0-1 Heavy	1-3 Medium	>3 Light

Data analyses

The data was analyzed using SPSS 17.0 (SPSS, 2008) and Microsoft Excel window 2007. Before analysis, the Kolmogorov-Smirnov method was used to test whether data were normally distributed, and the Bartlett test was performed to assess the homogeneity of variance. The data that did not meet the normality test was log transformed. One-way ANOVA was used to test the difference of the measures variable (environmental variables, phytoplankton diversity, richness and dominance) among sites. Relationships between phytoplankton species and environment variables were analyzed using the detrended correspondence analysis and redundancy analysis (RDA) using CANOCO 4.5 software (Microcomputer Power, New York, USA). Monte Carlo simulations with 499 permutations were used to test the significance of the physical–chemical variables in explaining the biomass of phytoplankton functional groups data in the RDA. Figures were drawn using Microsoft excel.

Results

Physical–chemical variables

The mean spatial values of physical–chemical factors recorded among the sampling sites within the study period are presented in Table 3. The average WT was 18.62 °C with the relative highest value of 20 °C being recorded in site 1#. The values of pH measured in this wetland park were in the side of alkalinity (basic). TN, TP, $\text{NH}_4^+\text{-N}$, $\text{NO}_3^-\text{-N}$ and CODcr differed significantly among the sites. TN, TP, $\text{NH}_4^+\text{-N}$ and $\text{NO}_3^-\text{-N}$ were significantly higher in site 2# with a mean value of 6.62, 0.67, 11.26 and 1.23 mg/L, respectively. Similarly, CODcr was significantly higher in site 2# (Table 2).

Classification which was done according to the *Environmental Quality Standard for Surface Water* (GB3838-2002), showed that with exception of $\text{NO}_3^- \text{N}$ which belongs to Class IV water quality, all the remaining measured variables belong to Class V water quality.

Table 3. Water quality measurement results of Taihu National Wetland Park (unit: except water temperature and pH, the rest are all mg/L)

Sampling site	WT(°C)	pH	$\text{NH}_4^+ \text{-N}$	$\text{NO}_3^- \text{-N}$	TN	TP	COD_{Cr}
1#	20	8.74	1.26	0.62	1.67	0.26	22.8
2#	18	8.25	11.26	1.23	6.62	0.67	29.6
3#	18	8.93	3.26	0.53	4.88	0.34	22.3
4#	19	9.17	2.21	0.59	4.82	0.27	20.2
5#	19	9.03	1.22	0.52	4.92	0.25	19.3
6#	19	9.32	0.96	0.39	3.46	0.18	2.6
7#	18	9.28	1.09	0.42	3.13	0.21	2.2
8#	18	9.11	1.06	0.46	2.95	0.18	2.3
Average value	18.62	8.98	2.79	0.59	4.06	0.30	15.2
GB3838-2002			V	IV	V	V	V

Phytoplankton species composition and their spatial distribution

In the current study, a total of 39 species of phytoplankton belonging to 27 genera and in 6 classes were identified. Chlorophyta was the most with 20 species, accounting for 51.28%; bacillariophyta was the second with 11 species accounting for 28.21%; Cyanophyta has 4 species, accounting for 10.26%; Euglenophyta has 2, accounting for 5.13%; Chrysophyta and cryptophyta had 1 each which accounted for 5.13% altogether.

Spatially, sampling site 5# had the highest number of species (31 species), followed by 6# (24 species) and the site with the least species number was 1# (Figure 2). The abundance and biomass of phytoplankton measured in the Taihu National Wetland Park was significantly different among sampling sites (Figure 3). There was spatial significant difference in abundance and biomass distribution among sampling sites (Figure 3). The abundance of the phytoplankton ranged between $141.6 \times 10^4 \text{ind./L} \sim 4311.6 \times 10^4 \text{ind./L}$, with the highest value of $4311.6 \times 10^4 \text{ind./L}$ recorded in site 5# while the least ($141.6 \times 10^4 \text{ind./L}$) was measured at site 1#. The phytoplankton biomass measured ranged between 0.39 mg / L and 21.24 mg / L, with the highest value of 21.24 mg / L measured at site 6# and the lowest of 0.39 mg / L at site 1#.

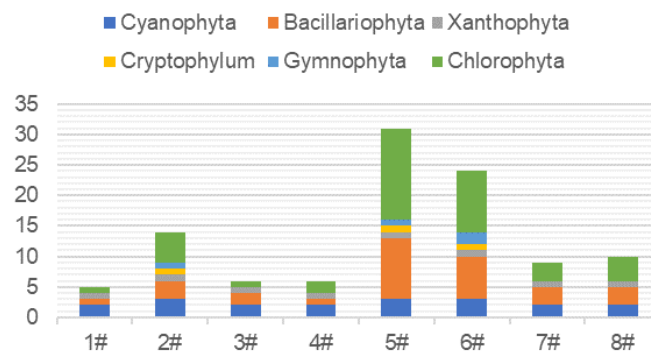


Figure 2. Number of phytoplankton species at sampling sites in Taihu National Wetland Park in autumn

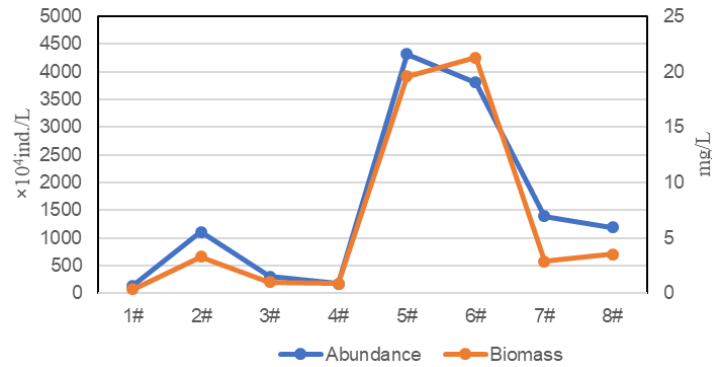


Figure 3. Phytoplankton abundance and biomass in Taihu National Wetland Park

Phytoplankton species diversity, richness and dominance

Phytoplankton community attributes such as diversity, richness and dominance often give important clues of the functional status or health of aquatic system. From the results, Shannon Weaver index ranged from 1.041 to 2.371 in the study area while; Margalef richness index ranged from 0.838 to 3.665 (Table 4 and Figure 4). Spatially, site 5# had the highest value of phytoplankton diversity of (2.371) while site 1# with a value of 1.041 was the least. Similarly, the taxon richness was significant higher in site 5# with a value of 3.665 and the least was observed in site 1# with a value of 0.838. Moreover, it is apparent that, Taihu National Wetland Park belongs to medium-heavy pollution type (Table 4).

Table 4. The diversity index and pollution evaluation of phytoplankton sampling points in Taihu National Wetland Park in autumn

Sampling site	<i>H'</i>	Assessment	<i>D</i>	Assessment
1#	1.041	Medium pollution	0.838	Heavy pollution
2#	1.248	Medium pollution	1.903	Medium pollution
3#	1.492	Medium pollution	0.901	Heavy pollution
4#	1.171	Medium pollution	1.001	Medium pollution
5#	2.371	Medium pollution	3.665	Light pollution
6#	2.229	Medium pollution	2.853	Medium pollution
7#	1.045	Medium pollution	1.134	Medium pollution
8#	1.287	Medium pollution	1.305	Medium pollution
Average value	1.486	Medium pollution	1.7	Medium pollution

Dominant species are defined by the dominance $y > 0.02$. The dominant species and their dominance of phytoplankton in Taihu National Wetland Park are shown in Table 5. In this study, the dominant species were *Schizosaccharum tenuiflorum*, filamentous alga bloom and *Scenedesmus giraldii*, belong to the indicator algae of eutrophic water body, and their dominance degrees are 0.675, 0.041 and 0.058, respectively.

When the occurrence of a species is greater than 65, then the species is regarded as a common species (Occurrence frequency > 65%). Table 6 shows the common species recorded during the study period in Taihu National Wetland Park. Seven common species belonging into 4 classes were observed in the wetland. These species are *Merismopedia minima*, *Aphanizomenon flosaquae*, *Cyclotella meneghiniana* and *Scenedesmus quadricauda*. These species are all indicator of eutrophic water body.

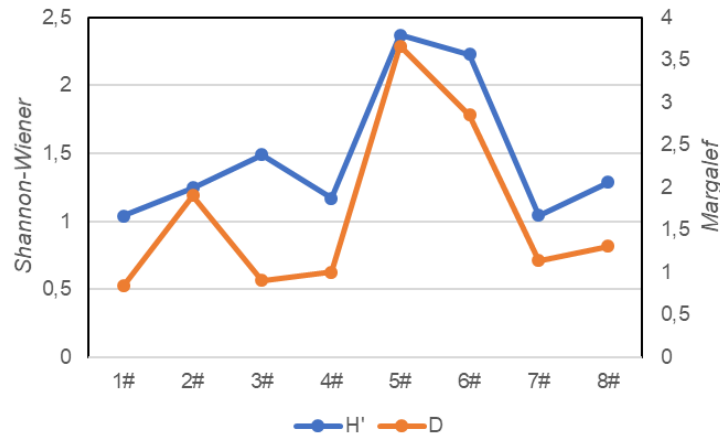


Figure 4. Phytoplankton taxon characteristics determined at the sites in the Taihu National Wetland Park

Table 5. The dominant species of phytoplankton in Taihu National Wetland Park in autumn

Dominant species	Occurrence frequency	Dominance
<i>Merismopedia minima</i>	1	0.675
<i>Aphanizomenon flosaquae</i>	1	0.041
<i>Synedra acus</i>	1	0.023
<i>Synura.sp.</i>	1	0.062
<i>Scenedesmus quadricauda</i>	1	0.058

Table 6. Common species of phytoplankton in autumn in Taihu National Wetland Park

Phylum	Species
Cyanophyta	<i>Merismopedia minima</i>
	<i>Aphanizomenon flosaquae</i>
Bacillariophyta	<i>Synedra acus</i>
	<i>Cyclotella meneghiniana</i>
Chrysophyta	<i>Synura.sp.</i>
Chlorophyta	<i>Ankistrodesmus angustus</i>
	<i>Scenedesmus quadricauda</i>

Relationship between environmental variables and some major phytoplankton species

RDA was carried out to determine the relationship between the abundance of the main functional species (Table 7) and environmental factors. The Monte Carlo test was significant for the first axis and all canonical axes ($p < 0.001$), suggesting that these environmental variables are important factors in explaining the group compositions. From Figure 5, $\text{NO}_3^- \text{-N}$ (0.8692) was the main positive correlation factor, TP (0.7523) was the second main positive correlation factor, and $\text{NH}_4^+ \text{-N}$ (0.6867) and COD_{Cr} (0.5641) also had significant positive correlation, while pH (-0.7595) in negative correlation factors had significant correlation. Water temperature (WT) was weakly but positive related to most of the phytoplankton species. From Figure 6, temperature and pH did not differ

significantly between the sites, therefore, these parameters could not influence the community structure.

Table 7. Species and numbers of phytoplankton in the RDA ordination chart

No.	Latin name	No	Latin name
L1	<i>Merismopedia minima</i>	Y1	<i>Cryptomonas ovate</i>
L2	<i>Merismopedia marssonii</i>	LU1	<i>Euglena oxyuris</i>
L3	<i>Phormidium allorgei</i>	LV1	<i>Chlamydomonas globosa</i>
L4	<i>Aphanizomenon flosaquae</i>	LV2	<i>Ankistrodesmus angustus</i>
G1	<i>Melosira granulata</i>	LV3	<i>Scenedesmus quadricauda</i>
G2	<i>Synedra acus</i>	LV4	<i>Selenastrum minutum</i> (Nag). Coll
G3	<i>Synedra amphicephala</i>	LV5	<i>Kirchneriella lunaris</i>
G4	<i>Synedra ulna</i>	LV6	<i>Dictyosphaerium pulchellum</i>
G5	<i>Cyclotella meneghiniana</i>	LV7	<i>Crucigenia tetrapedia</i>
G6	<i>Navicula exigua</i>	LV8	<i>Cosmarium obtusatum</i>
H1	<i>Synura.sp.</i>		

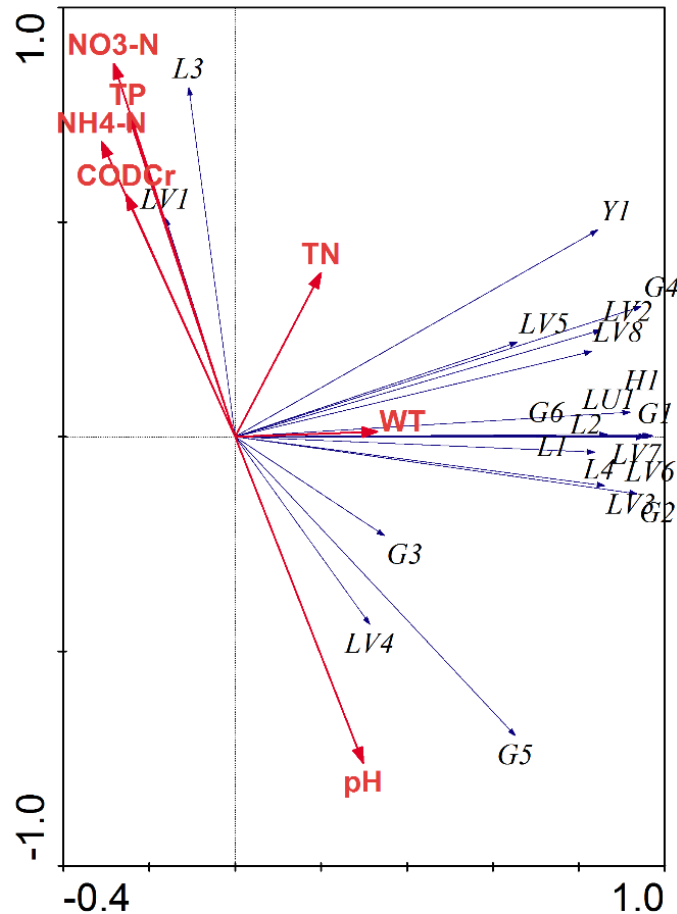


Figure 5. RDA analysis chart of phytoplankton species and environmental factors. Water temperature (WT), pH, total nitrogen (TN), total phosphorus (TP), ammonia nitrogen (NH_4^+ -N) and nitrate nitrogen (NO_3^- -N) and Chemical Oxygen Demand (CODCr)

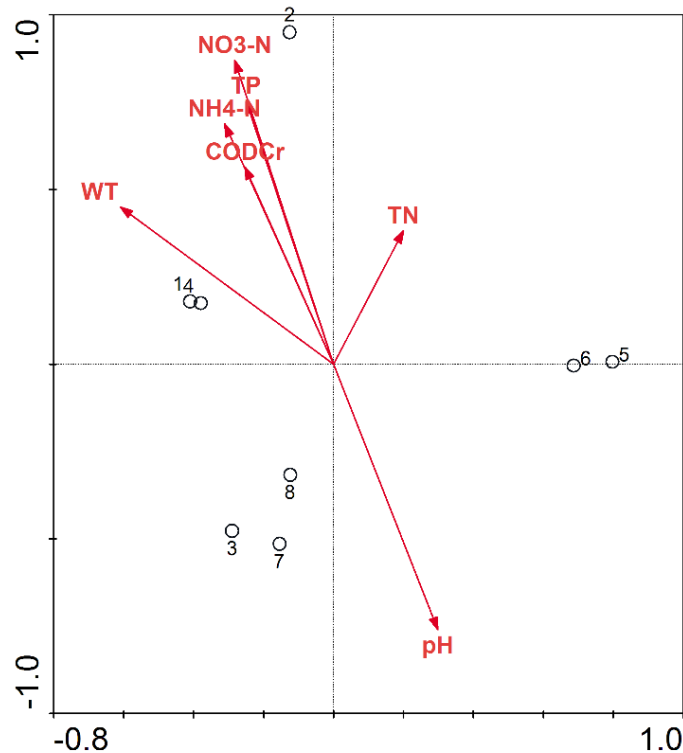


Figure 6. RDA analysis chart of sampling sites and environmental factors. Water temperature (WT), pH, total nitrogen (TN), total phosphorus (TP), ammonia nitrogen (NH_4^+ -N) and nitrate nitrogen (NO_3^- -N) and Chemical Oxygen Demand (CODCr).

Discussion

Variables such as dissolved gases, water temperature, pH, phosphates, nitrates, conductivity among many others including various physical properties (gases and solids solubility) are very important for growth and dispersal of phytoplankton (Mahar et al., 2000; Toma, 2011; Ma et al., 2019). Water temperature which is one of the important variables that influence growth of phytoplankton communities in aquatic system did not differ significantly. Water temperature not only influences the growth and distribution of phytoplankton in aquatic ecosystems, but also influences solubility of gases, water stratification, conductivity and pH (Sharm and Michael, 1987). The measured pH of lake varied between 8.25 to 9.32 during the study period. The values showed that the water of Taihu National Wetland Park was alkaline in nature.

Nutrients, especially phosphorus and nitrogen, are important for healthy growth of phytoplankton in aquatic systems. Plant can absorb nitrogen both as nitrate and ammonia. The concentrations of ammonia nitrogen, total nitrogen and total phosphorus went beyond class IV requirements of Environmental Quality Standards for Surface Water (National standards of the People's Republic of China) (GB3838-2002). The concentration data of total nitrogen and total phosphorus exceeded the limit values of eutrophication indicators proposed by OECD. Spatially, sites 2# and 3# had significant higher values of the nutrients concentration. This could be attributed by the fact that the site was located on a sewage treatment plant. Sites 5# to 8# recorded the lowest concentration values of nutrients probable because they were located within the wetland where the wetland plants and soil acted as filters of the nutrients.

In aquatic ecosystem, phytoplankton community structure often changes with the change of environment factors (Ouyangxb and Han, 2007; Li et al., 2014). Diversity indices has often been used to evaluate water pollution and phytoplankton community status. In this study, Shannon Weaver diversity index (H') was used to assess the diversity of phytoplankton spatial and also see whether environmental factors had an impact on the community structure. The study showed that the diversity index ranges from 1.041 to 2.371, with an average of 1.486. Species richness measured in this study ranged from 0.838 to 3.665, with an average of 1.7. Sites 5# and 6# which were light and medium polluted respectively had high diversity and richness values. According to the Intermediate Disturbance Hypothesis (IDH) theory, high diversity occurs at sites experiencing moderate levels of disturbance due to competition hierarchy of species. At low levels of disturbance or pollution, more competitive organisms will push subordinate species to extinction and dominate the ecosystem hence low diversity (Dial and Roughgarden, 1998). Those sites which were located far away from where the sewage treatment plant was had low phytoplankton diversity and richness. This implies that the distribution of phytoplankton species in Taihu National Wetland Park is uneven and the stability of community structure is unstable. Moreover, the fact that *Schizosaccharum tenuiflorum*, filamentous alga bloom and *Scenedesmus giraldii* were dominant implies that this wetland park could be facing serious pollution and also the concentration of organic matter is high (Liang et al., 2019).

The relationships between environmental variables and phytoplankton species were determined in this study. Water temperature (WT) was positively related to the abundance and biomass of the most phytoplankton species. According to Srifa (2010), water temperature is a factor that can positively or negatively affect the growth of some plankton species. Moreover, the rate of photosynthesis of phytoplankton is promoted by the rising water temperature, which accelerates the accumulation of biomass (Gong et al., 2020). The growth rate of phytoplankton is generally low in acidic water and higher in weak alkaline water (You et al., 2007). In this study, the pH value fluctuated from 8.25 to 9.32, which was beyond the pH value range of weak alkaline water, which might likely limit the growth and development of phytoplankton. Moreover, the results of RDA analysis showed that there was a significant negative correlation between pH and *Synedra amphicephala*, *Cyclotella meneghiniana* and *Selenastrum minutum* (Nag). Coll, which was consistent with the above phenomenon.

Nutrient enrichment will stimulate the growth of phytoplankton in relation to its abundance and biomass in various aquatic systems. Phytoplankton in different plant classes have different requirement for nutrition of phosphorus and nitrogen. Nitrogen species such as nitrate, nitrite and ammonia nitrogen, which are equilibrated with each other, can also affect phytoplankton community structure. The results of RDA analysis showed that there was a significant positive correlation between sampling site 2# and NO_3^- -N (Fig. 6), and the dominant species in this sampling site, *Phormidium allorgei* and *Chlamydomonas globosa*, also showed a significant positive correlation with NO_3^- -N and TP. This may be due to the fact that sampling site 2# is located in a sewage treatment plant with serious nitrogen and phosphorus pollution, and its ecological environment is conducive for the growth and development of *Phormidium allorgei* and *Chlamydomonas globosa*, which can be used as indicator species for mesotrophic-eutrophic water. Nutrients such as N and P are important for the growth of phytoplankton, and at particular concentrations can influence the growth of phytoplankton (Elmgren and Larsson, 2001). Sun et al. (2008) note that increase concentrations of N and P at a given range can promote

the growth of phytoplankton. This, therefore implies that nutrients factors in the wetland can effectively control the abundance and biomass of phytoplankton species through bottom-up effects.

Conclusion

In this study, a total of 39 species of phytoplankton were detected in Taihu National Wetland Park and Cyanophyta was dominant in some areas. The dominant species of phytoplankton in this survey are *Merismopedia minima* and *Aphanizomenon flosaquae* of Cyanophyta, *Synedra acus* of Bacillariophyta, *Synura.sp.* of Chrysophyta and *Scenedesmus quadricauda* of Chlorophyta. According to the evaluation of water quality based on physical and chemical indexes and phytoplankton diversity indexes, Taihu National Wetland Park is generally moderately to severely polluted, especially near the outlet of sewage treatment plant. Water temperature, nitrogen and phosphorus nutrients, and pH value are the key factors affecting phytoplankton community structure. Therefore, the above-mentioned environmental factors should be considered in the future treatment of water pollution in Taihu National Wetland Park.

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APPENDIX

Appendix 1. List of phytoplankton in Taihu National Wetland Park

Phylum	Family	Genus	Species	
Cyanophyta	Merismopediaceae	<i>Merismopedia</i>	<i>Merismopedia minima</i>	
			<i>Merismopedia marssonii</i>	
	Phormidioideae	<i>Phormidium</i>	<i>Phormidium allorgei</i>	
	Nostocaceae	<i>Aphanizomenon</i>	<i>Aphanizomenon flosaquae</i>	
Bacillariophyta	Fragilariaceae	<i>Fragilaria</i>	<i>Fragilaria brevistriata</i>	
		<i>Synedra</i>	<i>Synedra acus</i>	
			<i>Synedra amphicephala</i>	
			<i>Synedra ulna</i>	
		<i>Diatoma</i>	<i>Diatoma vulgare</i>	
	Coscinodiscaceae	<i>Melosira</i>	<i>Melosira granulata</i>	
		<i>Cyclotella</i>	<i>Cyclotella meneghiniana</i>	
	Naviculaceae	<i>Navicula</i>	<i>Navicula exigua</i>	
			<i>Navicula dicephala</i>	
	Gomphonemaceae	<i>Gomphonema</i>	<i>Gomphonema constrictum</i>	
	Cymbellaceae	<i>Cymbella</i>	<i>Cymbella ventricosa</i>	
Euglenophyta	Euglenaceae	<i>Euglena</i>	<i>Euglena oxyuris</i>	
		<i>Strombomonas</i>	<i>Strombomonas schauinslandii</i>	
Cyanophyta	Chlamydomonadaceae	<i>Chlamydomonas</i>	<i>Chlamydomonas globosa</i>	
	Chlorellaceae	<i>Ankistrodesmus</i>	<i>Ankistrodesmus angustus</i>	
			<i>Ankistrodesmus acicularis</i>	
			<i>Ankistrodesmus falcatus</i>	
		<i>Chodatella</i>	<i>Chodatella quadriseta</i>	
		<i>Selenastrum</i>	<i>Selenastrum gracile</i>	
			<i>Selenastrum minutum</i> (Nag). <i>coll</i>	
		<i>Kirchneriella</i>	<i>Kirchneriella lunaris</i>	
	Scenedesmaceae	<i>Scenedesmus</i>	<i>Scenedesmus bijuga</i>	
			<i>Scenedesmus dimorphus</i>	
			<i>Scenedesmus quadricauda</i>	
			<i>Scenedesmus platydiscus</i>	
			<i>Coelastrum</i>	<i>Coelastrum microporum</i>
		<i>Crucigenia</i>	<i>Crucigenia quadrata</i>	
			<i>Crucigenia tetrapedia</i>	
		Volvocaceae	<i>Pandorina</i>	<i>Pandorina morum</i>
		Pediastraceae	<i>Pediastrum</i>	<i>Pediastrum birtadiatum</i>
			<i>Pediastrum tetras</i>	
	Dictyosphaeraceae	<i>Dictyosphaerium</i>	<i>Dictyosphaerium pulchellum</i>	
	Desmidiaceae	<i>Cosmarium</i>	<i>Cosmarium obtusatum</i>	
Chrysophyta	Synuraceae	<i>Synura</i>	<i>Synura.sp</i>	
Cryptophyta	Cryptomonadaceae	<i>Cryptomonas</i>	<i>Cryptomonas ovata</i>	