

CAN LIFE HISTORIES PREDICT THE EFFECTS OF HABITAT FRAGMENTATION? A META-ANALYSIS WITH TERRESTRIAL MAMMALS

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Abstract. Anthropogenic fragmentation of habitats has been identified as one of the primary drivers of mammalian declines and extinctions. Previous research has implicated five life history traits as being predictive of the impacts of habitat fragmentation on mammalian abundances: potential growth rate, sociality, mass, home range, and niche breadth. In order to systematically test if these five life histories correlated with mammalian abundances across a gradient of habitat fragmentation, we conducted a meta-analysis. We systematically collected data from 68 studies, encompassing 232 mammalian species within 143 genera, 50 families, and 17 orders. We found that mammals with lower growth rates, paternal care of offspring, greater mass, larger home ranges, and increased niche specialization had significantly lower abundances in fragmented habitat. These results could provide land managers and conservationists with a coarse tool for predicting the impacts of habitat fragmentation across a wide taxonomic breadth of terrestrial mammals.

Keywords: *abundance, mixed-effects model, specialists vs. generalists*

Introduction

Humans have fragmented most continuous habitats on the planet, causing large changes in population sizes of many species (Turner, 1996; Ferraz et al., 2003). Anthropogenic fragmentation is the primary cause of mammalian declines and extinctions worldwide (Wilcove et al., 1998; Grelle, 2005), and has increasingly isolated parks and refuges created to protect wildlife (Janzen, 1983; Newmark, 1995). Despite their conservation status, these fragmented habitats frequently support fewer species, and the species that remain often maintain lower densities when compared to contiguous landscapes (Newmark, 1995).

Despite a large body of literature on the effects of fragmentation, it has been difficult to generalize results across studies because the research focus has typically been the *effects* of fragmentation on a species, and not *why* the species is affected by fragmentation (Funk & Mills, 2003; Banks et al., 2007). In other words, most studies addressed *how* a species is affected by fragmentation (for example, changes in demography or population size), and not what aspects of the species life-history caused the species to be vulnerable (e.g. Ceballos & Brown, 1995; Newmark, 1995; Brashares, 2002).

Proactive, strategic conservation necessitates an evidence-based framework that can

be used to predict the impact of anthropogenic disturbances. To accurately target conservation efforts, land managers need to know which species are most threatened by anthropogenic fragmentation and why (Doak & Mills, 1994; Turner, 1996). Several life-history traits have been proposed as predictive of the effects of fragmentation across species (e.g. Laurance, 1994; Ganzhorn & Eisenbeiß, 2001; Laurance et al., 2002), but the predictive capacity of these traits have not been systematically tested (Dale et al., 1994; Peters & Herrick, 2004; Banks et al., 2007), or applied systematically to predict impacts and aid management (Ferguson & Larivière, 2002; Funk & Mills, 2003).

To test whether selected life history traits predict mammal responses to fragmentation, we conducted a meta-analysis of all published literature linking fragmentation and mammals (Englund et al., 1999; Gurevitch & Hedges, 1999; Osenberg et al., 1999). Previous research has proposed that the potential growth rate, sociality, mass, home range, and niche breadth may each predict a directional change in mammal abundances due to fragmentation (*Table 1*). To assess whether the use of these life histories in management is warranted, we tested both the validity and strength of each of the following hypotheses.

Potential Growth Rate

Those species that have the greatest potential growth rate may be more capable of compensating for losses of individuals due to fragmentation (Laurance, 1991; Viveiros de Castro & Fernandez, 2004).

Sociality

Social mammals require groups in order to maintain populations or colonize fragments. Thus social mammals may be less likely to colonize a fragment and more prone to local extirpation than solitary species (Lawes et al., 2000; Swihart et al., 2003). Fragmentation can also have a negative effect on social species by reducing group size and therefore restricting fitness enhancing social strategies (Banks et al., 2007).

Mass

Mammals with heavier body mass require more resources. Because fragments provide a lower quantity of resources, mammals with large body masses may negatively correlate with fragmentation (Marquet & Taper, 1998; Cullen et al., 2001; Michalski & Peres, 2007; Okie & Brown, 2009).

Home Range

Wide ranging species are notably absent from small habitat fragments and may be particularly susceptible to habitat loss (Noss et al., 1996; Woodroffe & Ginsberg, 1998; Feeley & Terborgh, 2008). Thus as the home range of a species increases, the size of the fragment needed to support a viable population increases.

Niche Breadth

Specialists and generalists may respond differently to fragmentation. Specialists, compared to generalists, use fewer resources and thus have fewer alternatives when habitats are constricted. Specifically, specialization of diet, denning, and locomotion could predict a mammal's vulnerability to fragmentation (Laurance, 1990; Laurance,

1994; Swihart et al., 2003; Viveiros de Castro & Fernandez, 2004; Prugh et al., 2008).

Table 1. List of life histories and the predicted changes in mammal abundances in response to habitat fragmentation.

life history	effect of fragmentation	citations
low potential growth rate	negative	Laurance (1991) Viveiros de Castro & Fernandez (2004)
highly social	negative	Lawes et al. (2000) Swihart et al. (2003) Banks et al. (2007)
heavy mass	negative	Marquet & Taper (1998) Cullen et al. (2001) Michalski & Peres (2007) Okie & Brown (2009)
wide home range	negative	Noss et al. (1996) Woodroffe & Ginsberg (1998) Feeley & Terborgh (2008)
specialist	negative	Laurance (1990) Laurance (1994) Swihart et al. (2003) Viveiros de Castro & Fernandez (2004) Prugh et al. (2008)

Materials and Methods

Data Acquisition

On January 27th, 2010 we performed a literature search on Web of Science using three separate groups of phrases in order to avoid biases associated with selection criteria (Englund et al., 1999): 1) “fragment” and “mammal” and “patch,” 2) “fragment” and “mammal” and “habitat,” 3) “fragment” and “mammal” and “disturbance.” Because Web of Science detects word-fragments, a search for “fragment” returns all results for “fragmentation” as well. The results from these three groups of phrases were summed for a total of 1101 studies, and each study’s abstract was read. Based upon the abstracts, we read 304 studies in detail (see *Appendix 1* for the PRISMA flow diagram). Studies were selected for inclusion in the database if they assessed the abundances of terrestrial mammals within at least two unaltered fragments of different sizes. If a study had assessed abundances in fragments but did not include either the abundance estimates or fragment size, we contacted the author and requested the information for inclusion in this study. A total of 68 studies (*Appendix 2*) encompassing 232 mammalian species within 143 genera, 50 families, and 17 orders, were included in our meta-analysis.

Effect Statistic

We were interested in predicting shifts in mammalian abundances in response to anthropogenic fragmentation. To assess whether the aforementioned life history traits could predict changes in mammalian abundance, we needed to first quantify the impact

of fragmentation on abundances in the form of an effect statistic. In the traditional meta-analysis, the effect statistic is calculated as the standardized difference between two treatments (d in Hedges & Olkin, 1985) or the correlation coefficient (r in Osenberg et al., 1999). However, nearly all the studies collected for this meta-analysis assessed abundances in more than two habitat fragments; many species abundances were estimated in more than one study; and the goal of our research was to measure the magnitude of the effect of fragmentation on abundances; thus we needed an alternative statistic (Osenberg et al., 1997).

To accommodate these data, we used the regression coefficient from linear mixed-effects models as an effect statistic. By regressing the estimated abundances against the habitat fragment size within each species across studies, the resulting regression coefficient indicated both the strength and the direction of the effect for each species. We compared regression coefficients from linear and exponential regressions and found 127 of 203 species showed a better fit with a linear model. Given that the abundances of 85 species were estimated in two or more studies and given the need for one effect statistic per species in order to avoid pseudoreplication (Hurlbert, 1984), we included study as a random predictor with a fixed slope. Since both the fragment size and abundance estimates varied widely within and across studies, we used their respective logarithmic values (Michalski & Peres, 2007). In order to include abundance estimates of continuous forests in the model, we set their values equal to 10,000 ha (Vieira et al., 2009).

The effect statistic for each species is the negative of the regression coefficient ($-\beta_1$) from the following equation:

$$\log(\text{area}) \sim \beta_0 + \beta_1 * \log(\text{abundance}) + (1|\text{study}) \quad (\text{Eq. 1})$$

The (1|study) term allowed the various studies to have different intercepts, but also forced all the studies to have the same slope.

Life History Traits

To quantify a mammal's life history traits, we extracted values from encyclopedic references (Emmons & Feer, 1999; Nowak, 1999; Foresman & McGraw, 2001). If we could not find a given value, we searched for primary literature in Web of Science and in the IUCN Red List of Threatened Species (IUCN 2009). We defined potential growth rate as the litter size multiplied by the number of litters per year and then divided the total by generation time (Pianka, 1970). Each species' sociality received an ordinal score of 0 to 4, with asocial species scored as 0, rudimentary sociality as 1, matrilineal lineages but no male care as 2, monogamous species as 3, and communal groups with cooperative foraging as 4 (as in Swihart et al., 2003). Niche breadth was an ordinal value from 0 to 3 as calculated by the sum of a mammal's specialization across three axes, diet, denning, and locomotion. For each of these three traits, mammals received a 0 if they were a generalist and a 1 if they were a specialist. A generalist was defined as possessing an omnivorous diet, terrestrial locomotion, and terrestrial or fossorial denning (Laurance, 1994; Swihart et al., 2003; Viveiros de Castro & Fernandez, 2004; Prugh et al., 2008), and a specialist was anything else.

Statistical Analyses

To assess the predictive value of potential growth rate, mass, home range, and niche breadth, we used linear regressions. Potential growth rate, mass, and home range were continuous variables and niche breadth was an ordinal variable. Since land managers frequently do not know all five life histories for a given species, and since the goal of this research is to inform management, we regressed the effect statistic ($-\beta_1$ from Eq. 1) against each predictor individually. To assess possible inverse trends within order, we evaluated individual scatterplots with each taxonomic order. We also regressed the effect statistic against life histories for each taxonomic order represented by at least 10 species. Investigation of Cook's distance revealed fitted values with an influence greater than the 20th percentile for potential growth rate, mass, and home range. Thus we log-transformed these predictors (Kutner et al., 2005), and each of the regressions were as in the following equation:

$$\text{effect} \sim \beta_0 + \beta_1 * \text{life history} \quad (\text{Eq. 2})$$

We also compared ordinary linear models to linear mixed-effects models within which the individual slopes and intercepts were allowed to vary according to taxonomic order. Visual investigation of effect values as predicted by sociality revealed a discernible break point as differentiated by an absence (0, 1, 2) or presence (3, 4) of male care for offspring. To assess whether the presence or absence of male care predicted the effect of fragmentation on mammalian abundances, we performed a Wilcoxon rank sum test. All statistical analyses were run in R 2.10.1 (R Development Core Team, 2009).

Results

All five life history traits had a significant impact on the predicted effects of fragmentation on mammalian abundances. Higher potential growth rates reduced the impact of habitat fragmentation (*Table 2, Fig. 1a*; positive correlation with effect statistic), and species with paternal care were more negatively affected by fragmentation than those without paternal care (*Fig. 1b*; Wilcoxon rank sum, $W = 2947.5, p = 0.017$).

Table 2. Effect of fragmentation on mammalian abundances as predicted by life history traits. Below are results from the linear regression of the effect statistic against each predictor as in equation 2.

predictor	β_1	(\pm SE)	<i>p</i> -value	R ²
log (potential growth rate)	0.0209	(0.0062)	0.001	0.054
log (mass)	-0.0119	(0.0045)	0.009	0.029
log (home range)	-0.0100	(0.0043)	0.022	0.029
niche breadth	-0.0678	(0.0259)	0.009	0.030

Mass, home range, and niche breadth all showed significant negative correlations with the effect statistic (*Table 2*), indicating that species with heavier mass, larger home

range, and greater degree of specialization had lower abundances in habitat fragments (Fig. 1c, d, e).

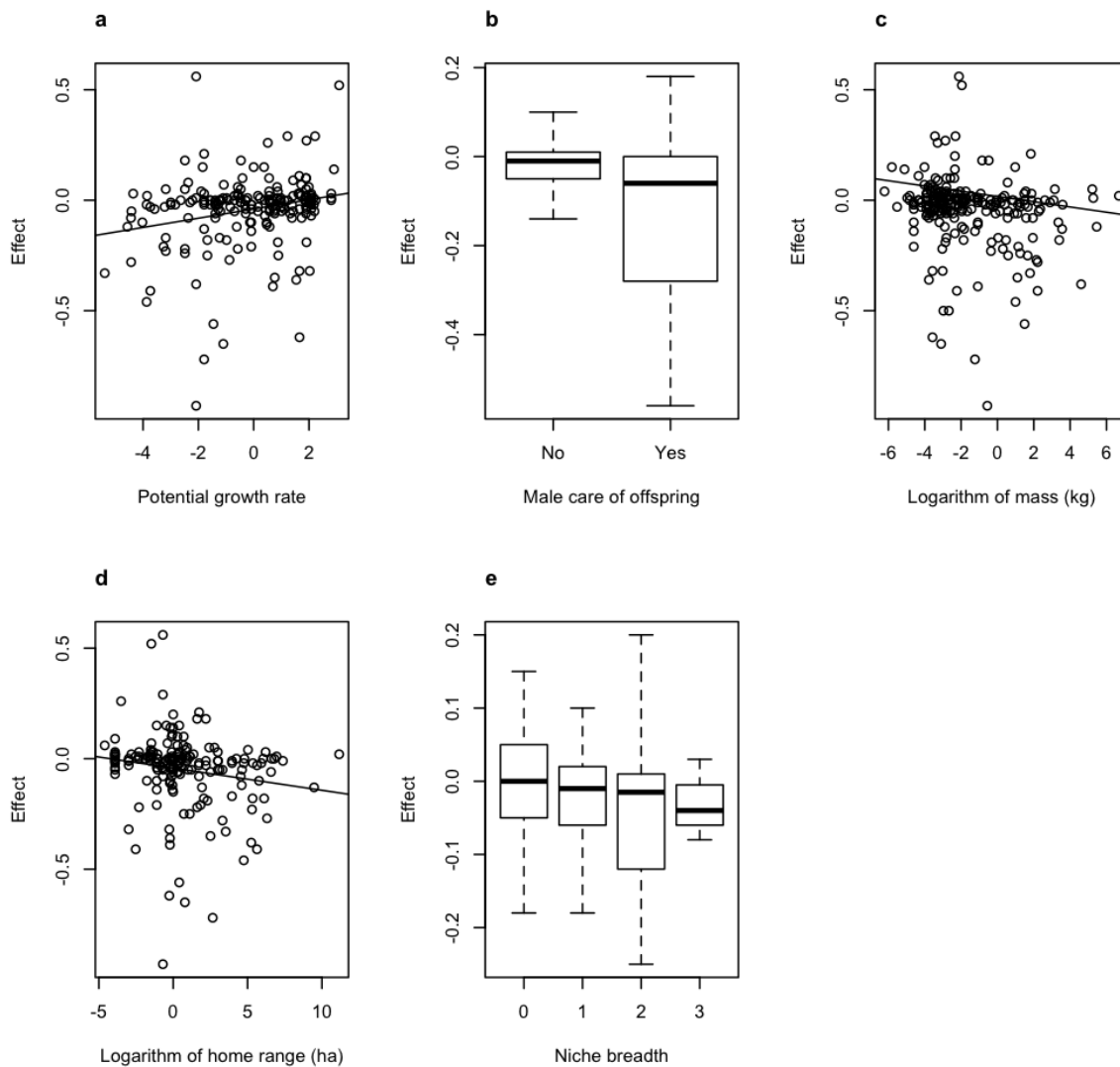


Figure 1. Effect of fragmentation on mammal abundances as predicted by (a) potential growth rate, (b) mammals with and without paternal care, (c) mass, (d) home range, and (e) increasing specialization across niche breadth.

The addition of taxonomic order as a random effect did not alter the interpretation of the results. With all five of the life histories, the coefficient and standard error for each regression changed less than 0.0003 with the addition of order as a random effect.

Linear regressions of the effect of fragmentation against life histories within taxonomic orders containing greater than 10 species revealed no significant trends (Table 3). We found no inverse trends within any of the five life history predictors indicating the trends found were not caused by over representation of any one order.

Discussion

In accordance with previous predictions, all five life history traits predicted mammalian abundances across a gradient of habitat fragmentation, and each of these correlations were significant. Thus the theorized tenets of mammalian life history as applied to the impacts of fragmentation appear sound, and generalizations found in a host of more specific studies appear to be substantiated by this meta-analysis.

Potential growth rate has been shown to correlate with Australian mammal abundances across a fragmented landscape, where species with lower growth rates had lower abundances in habitat fragments and vice versa (Laurance, 1991). Our research also indicates that mammals that are slow to mature and have few offspring are most likely to experience declines due to fragmentation (*Table 2, Fig. 1a*).

Social mammals can respond to fragmentation differently than asocial species. For example, the tree hyrax (*Dendrohyrax arboreaus*) and blue duiker (*Philantomba monticola*) do not exhibit parental care and had gradual declines due to fragmentation whereas the samango monkey (*Cercopithecus mitis*) has parental care and experienced dramatic declines in abundance (Lawes et al., 2000). The differences in abundances across these species were attributed to their social structure (Lawes et al., 2000). Our meta-analysis supports the observation that mammals with paternal care are more negatively affected by fragmentation than those without (*Fig. 1b*).

Several studies have found that larger mammals have lower abundances in habitat fragments than smaller species. In the Brazilian Atlantic forest, five common mammals of mass greater than 1 kg had average densities in 20,000 ha habitat fragments nearly triple that found in 200 ha fragments (Chiarello, 2000). On the Sunda shelf islands of Indonesia and Malaysia mammals of large body sizes were absent from smaller islands (Okie & Brown, 2009). Body size also accounts for carnivore abundances across coastal southern California with larger species having lower abundances in habitat fragments (Crooks, 2002). In the Chiquitano forests of Bolivia, mammals with a body mass less than 6 kg were found to have greater abundances in habitat fragments than in contiguous forests (Kosydar, 2010). Our results agree that mammals of greater mass are more susceptible to fragmentation (*Table 2, Fig. 1c*).

Mammals with larger home ranges need larger habitat fragments in order to support viable populations. Since people hunt mammals, the edges of habitat fragments can serve as a sink, thus rendering species with wide ranges especially susceptible to fragmentation (Woodroffe & Ginsberg, 1998). A study of ten carnivores found that mammals with large home ranges are more likely to go extinct than mammals with small home ranges (Woodroffe & Ginsberg, 1998). We found that this trend appears to hold true across taxa and that species with larger home ranges have lower abundances in fragmented habitats (*Table 2, Fig. 1d*).

The conservation literature points out differing responses of specialists and generalists to fragmentation. For example, a meta-analysis of amphibians, reptiles, invertebrates, birds and mammals in habitat fragments found that the specialization of diet and locomotion could predict the presence or absence of a species (Prugh et al., 2008). Results from the Biological Dynamics and Forest Fragments Project indicated that generalist herbivores and omnivores have stable or increasing abundances, whereas specialist predators have declined (Laurance et al., 2002). In Queensland, Australia specialization along the axes of diet, denning, and locomotion strongly correlated with abundances of five mammals with specialists nearly extirpated from habitat fragments (Laurance, 1990). The results of this meta-analysis indicate that resource specialization

in general – spanning all three of these axes – indicates a species response to fragmentation (Table 2, Fig. 1e).

Table 3. The effect of fragmentation as predicted by life histories within mammalian orders with at least 10 species represented.

order	potential growth rate		sociality		mass	
	β_1 (\pm SE)	<i>p</i> -value	W	<i>p</i> -value	β_1 (\pm SE)	<i>p</i> -value
Carnivora	-0.05 (0.06)	0.41	5	0.50	-0.01 (0.01)	0.93
Cetartiodactyla	-0.03 (0.05)	0.55	12	0.36	-0.01 (0.03)	0.80
Didelphimorphia	-0.04 (0.07)	0.59	no paternal care		-0.04 (0.02)	0.11
Diprotodontia	-0.05 (0.13)	0.69	10	1.00	0.04 (0.08)	0.66
Eulipotyphlya	0.02 (0.01)	0.10	no paternal care		-0.01 (0.01)	0.46
Primates	-0.01 (0.06)	0.82	11	0.43	0.00 (0.05)	0.92
Rodentia	0.01 (0.01)	0.24	308	0.54	-0.02 (0.01)	0.15

order	home range		niche breadth	
	β_1 (\pm SE)	<i>p</i> -value	β_1 (\pm SE)	<i>p</i> -value
Carnivora	0.01 (0.03)	0.80	0.06 (0.08)	0.50
Cetartiodactyla	0.00 (0.02)	0.96	-0.04 (0.10)	0.72
Didelphimorphia	-0.04 (0.02)	0.08	0.03 (0.07)	0.70
Diprotodontia	0.00 (0.15)	0.98	0.13 (0.20)	0.54
Eulipotyphlya	-0.01 (0.01)	0.67	0.01 (0.05)	0.81
Primates	-0.02 (0.04)	0.65	0.03 (0.09)	0.74
Rodentia	-0.01 (0.01)	0.60	-0.06 (0.04)	0.09

The above findings must be interpreted with care. Given that the trends we found were across taxonomic orders and not within taxonomic orders, we suggest that in the absence of additional research, these results should not be used to compare species within one order. Secondly, although we found a significant effect of fragmentation on abundances and we predicted this effect based upon a mammal's potential growth rate, sociality, mass, home range, and niche breadth, we also found a large degree of stochasticity with each of these predictors. Thus these results are best interpreted as suggestive of a mammal's potential response to fragmentation. Yet the stochasticity is not random. Given that we attempted to make global generalizations across a wide taxonomic array, a large amount variation should be expected.

Our findings indicate that the previously proposed theories generally hold and that each of the five life histories correlated with a mammal's response to fragmentation. For managers attempting to assess which mammals are most susceptible to habitat fragmentation, these life history traits can serve as a predictive tool. When developing plans about which species to monitor, this framework is likely to yield results allowing managers to focus limited resources on mammals that are prone to declines, specifically,

mammals with slow growth rates, paternal care, large mass, wide home ranges, and specialized niches. These five life history traits could provide a relative risk index that managers could use to help predict changes in abundances due to habitat fragmentation.

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REFERENCES

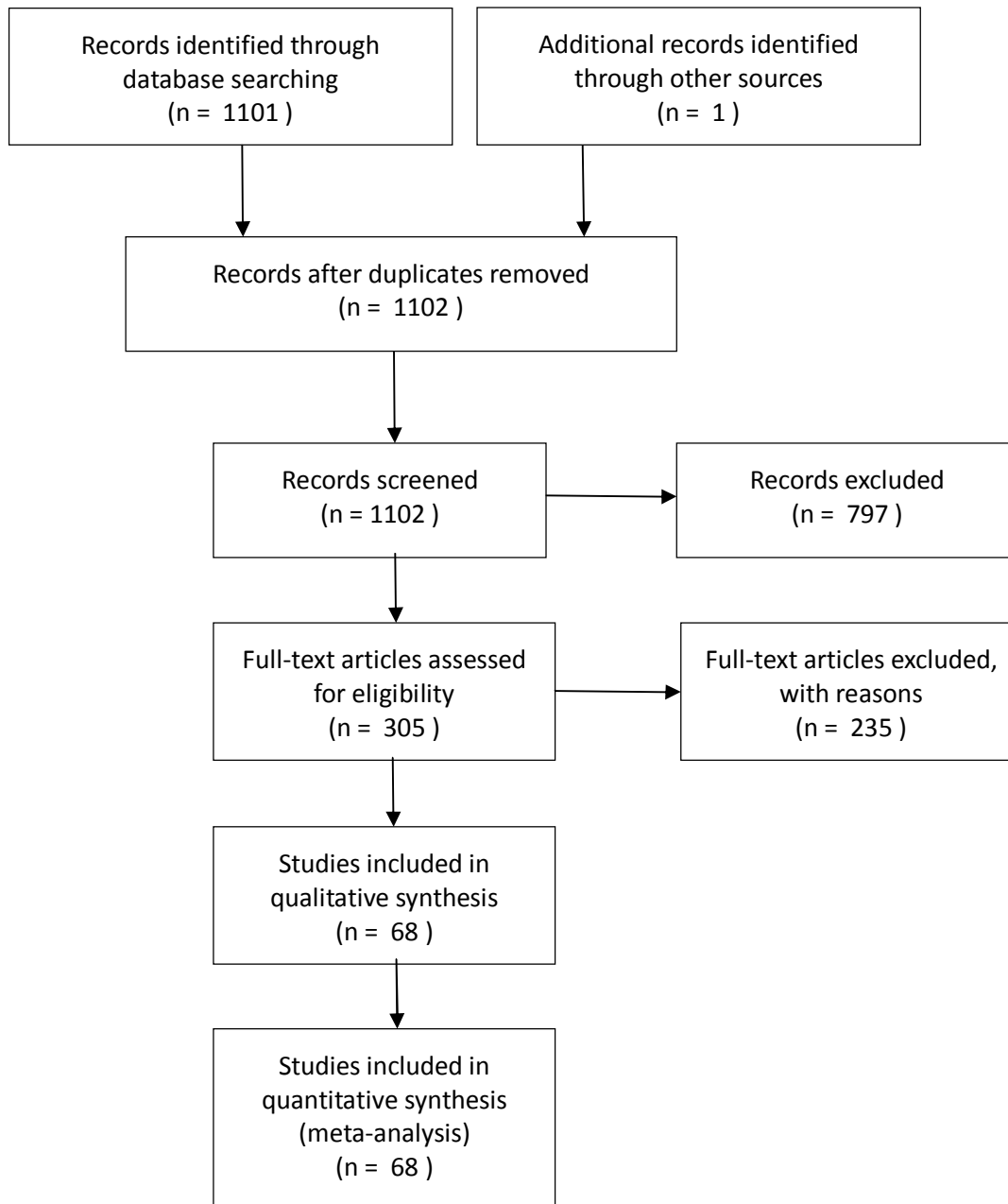
- [1] Banks, S.C., Piggott, M.P., Stow, A.J., Taylor, A.C. (2007): Sex and sociality in a disconnected world: a review of the impacts of habitat fragmentation on animal social interactions. – *Canadian Journal of Zoology* 85: 1065-1079.
- [2] Brashares, J. (2002): Ecological, behavioral, and life-history correlates of mammal extinctions in west Africa. – *Conservation Biology* 17: 733-743.
- [3] Ceballos, G., Brown, J.H. (1995): Global patterns of mammalian diversity, endemism, and endangerment. – *Conservation Biology* 9: 559-568.
- [4] Chiarello, A.G. (2000): Density and population size of mammals in remnants of Brazilian Atlantic forest. – *Conservation Biology* 14: 1649-1657.
- [5] Crooks, K.R. (2002): Relative sensitivities of mammalian carnivores to habitat fragmentation. – *Conservation Biology* 16: 488-502.
- [6] Cullen, L., Bodmer, E.R., Valladares-Padua, C. (2001): Ecological consequences of hunting in Atlantic forest patches, São Paulo, Brazil. – *Oryx* 35: 137-144.
- [7] Dale, V.H., Pearson, S.M., Offerman, H.L., O'Neill, R.O. (1994): Relating patterns of land-use change to faunal biodiversity in the central Amazon. – *Conservation Biology* 8: 1027-1036.
- [8] Doak, D.F., Mills, L.S. (1994): A useful role for theory in conservation. – *Ecology* 75: 615-626.
- [9] Emmons, L.H., Feer, F. (1999): Mamíferos de los bosques húmedos de América tropical: una guía de campo. – Fundación Amigos de la Naturaleza, Santa Cruz, Bolivia.
- [10] Englund, G., Sarnelle, O., Cooper, S.D. (1999): The importance of data-selection criteria: meta-analysis of stream predation experiments. – *Ecology* 80: 1132-1141.
- [11] Feeley, K.J., Terborgh, J.W. (2008): Trophic drivers of species loss from fragments. – *Animal Conservation* 11: 366-368.
- [12] Ferguson, S.H., Larivière, S. (2002): Can comparing life histories help conserve carnivores? – *Animal Conservation* 5: 1-12.
- [13] Ferraz, G., Russell, G.J., Stouffer, P.C., Bierregaard, R.O., Pimm, S.L., Lovejoy, T.E. (2003): Rates of species loss from Amazonian forest fragments. – *Proceedings of the National Academy of Sciences* 100: 14069-14073.
- [14] Foresman, K.R., McGraw, R.L. (2001): The wild mammals of Montana. – American Society of Mammalogists, Lawrence, Kansas.
- [15] Funk, C.F., Mills, L.S. (2003): Potential causes of population declines in forest fragments in an Amazonian frog. – *Biological Conservation* 111: 205-214.
- [16] Ganzhorn, J.U., Eisenbeiß, B. (2001): The concept of nested species assemblages and its utility for understanding effects of habitat fragmentation. – *Basic and Applied Ecology* 2: 87-95.
- [17] Grelle, C.E.V. (2005): Predicting extinction of mammals in the Brazilian Amazon. – *Oryx* 39: 347-350.
- [18] Gurevitch, J., Hedges, L.V. (1999): Statistical issues in ecological meta-analyses. –

- Ecology 80: 1142-1149.
- [19] Hedges, L.V., Olkin, I. (1985): Statistical methods for meta-analysis. – Academic Press, Orlando, Florida.
- [20] Hurlbert, S.H. (1984): Pseudoreplication and the design of ecological field experiments. – Ecological Monographs 54: 187-211.
- [21] IUCN (2009): IUCN red list of threatened species. – Version 2009.2. www.iucnredlist.org
- [22] Janzen, D.H. (1983): No park is an island: increase in interference from outside as park size decreases. – Oikos 41: 402-410.
- [23] Kosydar, A. (2010): Predicting the unpredictable: mammals in fragmented landscapes. – PhD dissertation. Department of Biology, University of Washington. Seattle, USA.
- [24] Kutner, M.H., Nachtsheim, C.J., Neter, J., Li, W. (2005): Applied linear statistical models. 5th edition. – McGraw-Hill, Boston.
- [25] Laurance, W.F. (1990): Comparative responses of five arboreal marsupials to tropical forest fragmentation. – Journal of Mammalogy 71: 641-653.
- [26] Laurance, W.F. (1991): Ecological correlates of extinction proneness in Australian tropical rain forest mammals. – Conservation Biology 5: 79-89.
- [27] Laurance, W.F. (1994): Rainforest fragmentation and the structure of small mammal communities in tropical Queensland. – Biological Conservation 69: 23-32.
- [28] Laurance, W.F., Lovejoy, T.E., Vasconcelos, H.L., Bruna, E.M., Didham, R.K., Stouffer, P.C., Gascon, C., Bierregaard, R.O., Laurance, S.G., Sampaio, E. (2002): Ecosystem decay of Amazonian forest fragments: a 22-year investigation. – Conservation Biology 16: 605-618.
- [29] Lawes, M.J., Mealin, P.E., Piper, S.E. (2000): Dynamics of three forest mammals in fragmented afro-montane forest in South Africa. – Conservation Biology 14: 1088-1098.
- [30] Marquet, P.A., Taper, M.L. (1998): On size and area: patterns of mammalian body size extremes across landscapes. – Evolutionary Ecology 12: 127-139.
- [31] Michalski, F., Peres, C.A. (2007): Disturbance-mediated mammal persistence and abundance-area relationships in Amazonian forest fragments. – Conservation Biology 21: 1626-1640.
- [32] Moher, D., Liberati, A., Tetzlaff, J., Altman, D.G., The PRISMA Group (2009): Preferred Reporting Items for Systematic Reviews and Meta-Analyses: The PRISMA Statement. – PLoS Med 6: e1000097. doi:10.1371/journal.pmed1000097
- [33] Newmark, W.D. (1995): Extinction of mammal populations in Western North American national parks. – Conservation Biology 9: 512-526.
- [34] Noss, R.F., Quigley, H.B., Hornocker, M.G., Merrill, T., Paquet, P.C. (1996): Conservation biology and carnivore conservation in the Rocky mountains. – Conservation Biology 10: 949-963.
- [35] Nowak, R.M. (1999): Walker's Mammals of the World. Sixth Edition. Volume I. – Johns Hopkins University Press, Baltimore and London.
- [36] Okie, J.G., Brown, J.H. (2009): Niches, body sizes, and the disassembly of mammal communities on the Sunda Shelf islands. – Proceedings of the National Academy of Sciences 106: 19679-19684.
- [37] Osenberg, C.W., Sarnelle, O., Cooper, S.D. (1997): Effect size in ecological experiments: the application of biological models in meta-analysis. – The American Naturalist 150: 798-812.
- [38] Osenberg, C.W., Sarnelle, O., Cooper, S.D., Holt, R.D. (1999): Resolving ecological questions through meta-analysis: goals, metrics, and models. – Ecology 80: 1105-1117.
- [39] Pianka, L.F. (1970): On *r*- and *K*-selection. – American Naturalist 104: 592-597.
- [40] Peters, D.P.C., Herrick, J.E. (2004): Strategies for ecological extrapolation. – Oikos 106: 627-636.
- [41] Prugh, L.R., Hodges, K.E., Sinclair, A.R.E., Brashares, J.S. (2008): Effect of habitat area and isolation on fragmented animal populations. – Proceedings of the National Academy

- of Sciences 105: 20770-20775.
- [42] R Development Core Team (2009): R: a language and environment for statistical computing. – R Foundation for Statistical Computing, Vienna, Austria. www.R-project.org
- [43] Swihart, R.K., Atwood, T.C., Goheen, J.R., Scheiman, D.M., Munroe, K.E., Gehring, T.M. (2003): Patch occupancy of North American mammals: is patchiness in the eye of the beholder? – *Journal of Biogeography* 30: 1259-1279.
- [44] Turner, I.M. (1996): Species loss in fragments of tropical rain forest: a review of the evidence. – *Journal of Applied Ecology* 33: 200-209.
- [45] Vieira, M.V., Olifiers, N., Delciellos, A.C., Antunes, V.Z., Bernardo, L.R., Grelle, C.E.V., Cerqueira, R. (2009): Land use vs. fragment size and isolation as determinants of small mammal composition and richness in Atlantic forest remnants. – *Biological Conservation* 142: 1191-1200.
- [46] Viveiros de Castro, E.B., Fernandez, F.A.S. (2004): Determinants of differential extinction vulnerabilities of small mammals in Atlantic forest fragments in Brazil. – *Biological Conservation* 119: 73-80.
- [47] Wilcove, D.S., Rothstein, D., Dubow, J., Phillips, A., Losos, E. (1998): Quantifying threats to imperiled species in the United States. – *BioScience* 48: 607-615.
- [48] Woodroffe, R., Ginsberg, J.R. (1998): Edge effects and the extinction of populations inside protected areas. – *Science* 280: 2126-2128.

Appendices

Appendix 1. PRISMA Flow Diagram (Moher et al., 2009) detailing the flow of information through the different phases of the meta-analysis. This flow diagram maps out the number of records identified, included and excluded, and the reasons for exclusions.



Appendix 2. List of 68 studies from which data was extracted for the meta-analysis. An asterisk (*) in front of the title indicates that the author(s) of the study shared additional data for inclusion in this meta-analysis.

1. Anderson, C.S., Cady, A.B., Meikle, D.B. (2003): Effects of vegetation structure and edge habitat on the density and distribution of white-footed mice (*Peromyscus leucopus*) in small and large forest patches. – Canadian Journal of Zoology-Revue Canadienne de Zoologie 81: 897–904.
2. * Barko, V.A., Feldhamer, G.A., Nicholson, M.C., Davie, D.K. (2003): Urban habitat: a determinant of white-footed mouse (*Peromyscus leucopus*) abundance in southern Illinois. – Southeastern Naturalist 2: 369–376.
3. Bayne, E.M., Hobson, K.A. (1998): The effects of habitat fragmentation by forestry and agriculture on the abundance of small mammals in the southern boreal mixed wood forest. – Canadian Journal of Zoology-Revue Canadienne de Zoologie 76: 62–69.
4. Bentley, J.M., Catterall, C.P. Smith, G.C. (2000): Effects of fragmentation of Araucarian vine forest on small mammal communities. – Conservation Biology 14: 1075–1087.
5. Chiarello, A.G. (2000): Density and population size of mammals in remnants of Brazilian Atlantic forest. – Conservation Biology 14: 1649–1657.
6. Chiarello, A.G., De Melo, F.R. (2001): Primate population densities and sizes in Atlantic forest remnants of northern Espirito Santo, Brazil. – International Journal of Primatology 22: 379–396.
7. Cote, M., Ferron, J. (2001): Short-term use of different residual forest structures by three sciurid species in a clear cut boreal landscape. – Canadian Journal of Forest Research-Revue Canadienne De Recherche Forestiere 31: 1805–1815.
8. Cox, M.P., Dickman, C.R., Hunter, J. (2004): Effects of rainforest fragmentation on non-flying mammals of the eastern Dorrigo plateau, Australia. – Biological Conservation 115: 175–189.
9. Cullen, L.J. (1997): Hunting and biodiversity in Atlantic forest fragments, Sao Paulo, Brazil. – MS Thesis. University of Florida. Gainesville, USA.
10. De Araujo, R.M., de Souza, M.B., Ruiz-Miranda, C.R. (2008): Density and population size of game mammals in two Conservation Units of the State of Rio de Janeiro, Brazil. – Iheringia Serie Zoologia 98: 391–396.
11. Diaz, M., Santos, T., Telleria, J.L. (1999): Effects of forest fragmentation on the winter body condition and population parameters of a habitat generalist, the wood mouse *Apodemus sylvaticus*: a Test of Hypotheses. – Acta Oecologica-International Journal of Ecology 20: 39–49.
12. Dunstan, C.E., Fox, B.J. (1996): The effects of fragmentation and disturbance of rainforest on ground-dwelling small mammals on the Robertson plateau, New South Wales, Australia. – Journal of Biogeography 23: 187–201.
13. Elliott, A.G., Root, B.G. (2006): Small mammal responses to silvicultural and precipitation-related disturbance in northeastern Missouri riparian forests. – Wildlife Society Bulletin 34: 485–501.
14. Fisher, J.T., Merriam, G. (2000): Resource patch array use by two squirrel species in an agricultural landscape. – Landscape Ecology 15: 333–338.
15. Ganzhorn, J.U. (2003): Effects of introduced *Rattus rattus* on endemic small mammals in dry deciduous forest fragments of western Madagascar. – Animal Conservation 6: 147–157.

16. Gottfried, B.M. (1997): Small mammal populations in woodlot islands. – *American Midland Naturalist* 102: 105–112.
17. Granjon, L., Cosson, J.F., Judas, J., Ringuet, S. (1996): Influence of tropical rainforest fragmentation on mammal communities in French Guiana: short-term effects. – *Acta Oecologica-International Journal of Ecology* 17: 673–684.
18. Hanser, S.E., Huntly, N.J. (2006): The biogeography of small mammals of fragmented sagebrush-steppe landscapes. – *Journal of Mammalogy* 87: 1165–1174.
19. Hanski, I. (1986): Shrews on small islands: epigenetic variation elucidates population stability. – *Holarctic Ecology* 9: 193–204.
20. Happold, D.C.D., Happold, M. (1997): Conservation of mammals on a tobacco farm on the highlands of Malawi. – *Biodiversity and Conservation* 6: 837–852.
21. Harrington, G.N., Freeman, A.N.D., Crome, F.H.J. (2001): The effects of fragmentation of an Australian tropical rain forest on populations and assemblages of small mammals. – *Journal of Tropical Ecology* 17: 225–240.
22. Hayward, M.W., De Tores, P.J., Dillon, M.J., Fox, B.J. (2003): Local population structure of a naturally occurring metapopulation of the Quokka (*Setonix brachyurus* Macropodidae : Marsupialia). – *Biological Conservation* 110: 343–355.
23. Henderson M.T., Merriam, G., Wegner, J. (1985): Patchy environments and species survival: chipmunks in an agricultural mosaic. – *Biological Conservation* 31: 95–105.
24. Jackson, S.M. (2000): Population dynamics and life history of the mahogany glider, *Petaurus gracilis*, and the sugar glider, *Petaurus breviceps*, in north Queensland. – *Wildlife Research* 27: 21–37.
25. Johnson, R., Ferguson, J.W.H., Van Jaarsveld, A.S., Bronner, G.N., Chimimba, C.T. (2002): Delayed responses of small-mammal assemblages subject to afforestation-induced grassland fragmentation. – *Journal of Mammalogy* 83: 290–300.
26. Kelt, D.A. (2000): Small mammal communities in rainforest fragments in central southern Chile. – *Biological Conservation* 92: 345–358.
27. Knight, E.H., Fox, B.J. (2000): Does habitat structure mediate the effects of forest fragmentation and human-induced disturbance on the abundance of *Antechinus stuartii*? – *Australian Journal of Zoology* 48: 577–595.
28. Kosydar, A. (2010): Predicting the unpredictable: mammals in fragmented landscapes. – PhD dissertation. Department of Biology, University of Washington. Seattle, USA.
29. Krohne, D.T., Hoch, G.A. (1999): Demography of *Peromyscus Leucopus* Populations on Habitat Patches: the Role of Dispersal. – *Canadian Journal of Zoology-Revue Canadienne de Zoologie* 77: 1247–1253.
30. Laakkonen, J., Fisher, R.N., Case, T.J. (2001): Effect of land cover, habitat fragmentation and ant colonies on the distribution and abundance of shrews in southern California. – *Journal of Animal Ecology* 70: 776–788.
31. Lambert, T.D., Adler, G.H., Riveros, C.M., Lopez, L., Ascanio, R., Terborgh, J. (2003): Rodents on tropical land-bridge islands. – *Journal of Zoology* 260: 179–187.
32. Laurance, W.F. (1990): Comparative responses of five arboreal marsupials to tropical forest fragmentation. – *Journal of Mammalogy* 71: 641–653.
33. Laurance, W.F. (1994): Rainforest fragmentation and the structure of small mammal communities in tropical Queensland. – *Biological Conservation* 69: 23–32.
34. Leung L.K.P., Dickman, C.R., Moore, L.A. (1993): Genetic variation in fragmented populations of an Australian rainforest rodent, *Melomys cervinipes*. – *Pacific*

- Conservation Biology 1: 58–65.
35. * Loman, J. (1991): Small mammal and raptor densities in habitat islands; area effects in a south Swedish agricultural landscape. – *Landscape Ecology* 5: 183–189.
 36. Lynam, A.J., Billick, I. (1999): Differential responses of small mammals to fragmentation in a Thailand tropical forest. – *Biological Conservation* 91: 191–200.
 37. Malcolm, J.R. (1988): Small mammal abundances in isolated and non-isolated primary forest reserves near Manaus, Brazil. – *Acta Amazonica* 18: 67–83.
 38. Marchesan, D., Carthew, S.M. (2004): Autecology of the yellow-footed *Antechinus* (*Antechinus flavipes*) in a fragmented landscape in southern Australia. – *Wildlife Research* 31: 273–282.
 39. Marsh, A.C.W., Harris, S. (2000): Partitioning of woodland habitat resources by two sympatric species of *Apodemus*: lessons for the conservation of the yellow-necked mouse (*A. flavicollis*) in Britain. – *Biological Conservation* 92: 275–283.
 40. Matthiae P.E., Sterns, F. (1981): Mammals in forest islands in southeastern Wisconsin. – In: Burgess, R.L., Sharde, D.M. (eds.) *Forest island dynamics in man-dominated landscapes*. Springer, New York.
 41. Michalski, F., Peres, C.A. (2007): Disturbance-mediated mammal persistence and abundance-area relationships in Amazonian forest fragments. – *Conservation Biology* 21: 1626–1640.
 42. Mills, L.S. (1995): Edge effects and isolation: red-backed voles on forest remnants. – *Conservation Biology* 9: 395–402.
 43. * Mudappa, D., Kumar, A., Chellam, R. (2001): Abundance and habitat selection of the Malabar spiny dormouse in the rainforests of the southern western Ghats, India. – *Current Science* 80: 424–427.
 44. Nakagawa, M., Miguchi, H., Nakashizuka, T. (2006): The effects of various forest uses on small mammal communities in Sarawak, Malaysia. – *Forest Ecology and Management* 231: 55–62.
 45. Norris, D., Peres, C.A., Michalski, F., Hinchsliffe, K. (2008): Terrestrial mammal responses to edges in Amazonian forest patches: a study based on track stations. – *Mammalia* 72: 15–23.
 46. * Nupp, T.E., Swihart, R.K. (2000): Landscape level correlates of small mammal assemblages in forest fragments of farmland. – *Journal of Mammalogy* 81: 512–526.
 47. * Oguge, N., Hutterer, R., Odhiambo, R., Verheyen, W. (2004): Diversity and structure of shrew communities in montane forests of southeast Kenya. – *Mammalian Biology* 69: 289–301.
 48. Pahl Li, Winter, J., Heinsohn, G. (1988): Variation in responses of arboreal marsupials to fragmentation of tropical rainforest in north eastern Australis. – *Biological Conservation* 46: 71–82.
 49. Pardini, R., de Souza, S. M., Braga-Neto, R. & Metzger, J. P. (2005): The role of forest structure, fragment size and corridors in maintaining small mammal abundance and diversity in an Atlantic forest landscape. – *Biological Conservation* 124: 253–266.
 50. Pattanavibool, A. (1999): Wildlife response to habitat fragmentation and other human influences in tropical montane evergreen forests, northern Thailand. – PhD dissertation. University of Victoria. Victoria, Canada.
 51. Pires, A.S., Lira, P.K., Fernandez, F.A.S., Schittini, G.M., Oliveira, L.C. (2002): Frequency of movements of small mammals among Atlantic coastal forest fragments in Brazil. – *Biological Conservation* 108: 229–237.

52. * Price, O., Rankmore, B., Milne, D., Brock, C., Tynan, C., Kean, L., Roeger, L. (2005): Regional patterns of mammal abundance and their relationship to landscape variables in eucalypt woodlands near Darwin, northern Australia. – *Wildlife Research* 32: 435–446.
53. Quental, T. B. Fernandez, F.A.D.S., Dias, A.T.C., Rocha, F.S. (2001): Population dynamics of the marsupial *Micoureus demerarae* in small fragments of Atlantic coastal forest in Brazil. – *Journal of Tropical Ecology* 17: 339–352.
54. Ramanamanjato, J.B., Ganzhorn, J.U. (2001): Effects of forest fragmentation, introduced *Rattus rattus* and the role of exotic tree plantations and secondary vegetation for the conservation of an endemic rodent and a small lemur in littoral forests of southeastern Madagascar. – *Animal Conservation* 4: 175–183.
55. Redfield, J. A. (1976): Distribution, abundance, size, and genetic variation of *Peromyscus maniculatus* on the Gulf Islands of British Columbia. – *Canadian Journal of Zoology-Revue Canadienne de Zoologie* 54: 463–474.
56. Saavedra, B., Simonetti, J.A. (2005): Small mammals of Maulino forest remnants, a vanishing ecosystem of south central Chile. – *Mammalia* 69: 1–12.
57. Santos T, Telleria, J.L., Virgós, E. (1999) Dispersal of Spanish juniper *Juniperus thurifera* by birds and mammals in a fragmented landscape. – *Ecography* 22: 193–204.
58. Shanker, K., Sukumar, R. (1998): Community structure and demography of small mammal populations in insular montane forests in southern India. – *Oecologia* 116: 243–251.
59. Silva, M. (2001): Abundance, diversity, and community structure of small mammals in forest fragments in Prince Edward Island National Park, Canada. – *Canadian Journal of Zoology-Revue Canadienne de Zoologie* 79: 2063–2071.
60. Smith A.T., Vrieze, J.M. (1978) Population structure of Everglades rodents: responses to a patchy environment. – *Journal of Mammalogy* 60: 778-794.
61. Sridhar, H., Raman, T.R.S., Mudappa, D. (2008): Mammal persistence and abundance in tropical rainforest remnants in the southern western Ghats, India. – *Current Science* 94: 748–757.
62. Stevens, S.M., Husband, T.P. (1998): The influence of edge on small mammals: evidence from Brazilian Atlantic forest fragments. – *Biological Conservation* 85: 1–8.
63. Suchomel, J., Heroldova, M. (2004): Small terrestrial mammals in two types of forest complexes in intensively managed landscape of south Moravia (the Czech Republic). – *Ekologia-Bratislava* 23: 377–384.
64. Telleria, J.L., Santos, T., Alcantara, M. (1991): Abundance and food searching intensity of wood mice (*Apodemus sylvaticus*) in fragmented forests. – *Journal of Mammalogy* 72: 183–187.
65. van Alperdoorn, R.C., Oostenbrink, W.T., van Winden, A., van der Zee, F.F. (1992): Effect of habitat fragmentation on the bank vole, *Clethrionomys glareolus*, in an agricultural landscape. – *Oikos* 65: 265-274.
66. Vieira, M., Olifiers, N., Delciellos, A.C., Antunes, V.Z., Bernardo, L.R., Grelle, C.E.V., Cerqueira, R. (2009): Land use vs. fragment size and isolation as determinants of small mammal composition and richness in Atlantic Forest remnants. – *Biological Conservation* 142: 1191–1200.
67. Wilder, S.M., Abtahi, A.M., Meikle, D.B. (2005): The effects of forest fragmentation on densities of white-footed mice (*Peromyscus leucopus*) during the winter. –

- American Midland Naturalist 153: 71–79.
68. Witt, W.C., Huntly, N. (2001): Effects of isolation on red-backed voles (*Clethrionomys gapperi*) and deer mice (*Peromyscus maniculatus*) in a sage-steppe matrix. – Canadian Journal of Zoology-Revue Canadienne de Zoologie 79: 1597–1603.