FLAME ATOMIC ABSORPTION SPECTROMETRY DETERMINATION OF SILVER NANOPARTICLES IN ENVIRONMENTAL WATERS USING DISPERSIVE LIQUID-LIQUID MICROEXTRACTION

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Abstract. A novel, rapid low-density solvent based dispersive liquid-liquid microextraction has been developed for extraction and preconcentration of silver nanoparticles (AgNPs) prior to flame atomic absorption spectrometry analysis. In this study, water samples were first treated with acetonitrile/water (80:20, v/v) which was then utilized as disperser solvent in the dispersive liquid-liquid microextraction process along with using 1-octanol as extractant. An enrichment factor of 57 is obtained, and the recoveries of AgNPs spiked into environmental samples are in the range of 96.7 to 108%. A low limit of detection (LOD) of 6.5 µg/L is achieved using flame atomic absorption spectrometry (FAAS) for quantification. The relative standard deviation (RSD) is 2.1% (n = 9). This proposed method provides an efficient approach for the analysis and tracking of AgNPs in the environmental water samples. **Keywords:** *AgNPs*, *DLLME*, *extractant*, *preconcentration*, *FAAS*

Introduction

Metallic nanoparticles are of high importance in current materials research, with potential applications in biomedicine, energy conversion, imaging, or as pigments (Khaligh et al., 2015; Hartmann et al., 2013). Engineered nanomaterials (NMs) have already yielded a variety of commercially available products including cosmetics, antimicrobials, suntan lotions, paints, stain-resistant clothing and remediation products (Mirzaei et al., 2013). The rapid growth in the commercial use of AgNPs will inevitably increase silver exposure to the general population (Ghaedi et al., 2013). AgNPs may penetrate the skin (Ghaedi et al., 2015) via textile or wound dressing contact, be inhaled from some spray medicine, or ingress into the female genital tract by incorporation into female hygienic products or medical implants (Habila et al., 2015), and it is also toxic to human cells (Altunay et al., 2015; Fan et al., 2012; Choi et al., 2011). Therefore, it is important from an analytical point of view to develop sensitive and economical methods for determination of trace silver nanoparticles.

Because of the complexity of the matrices and trace amounts of metal ions residues in practical samples, sample preparation is necessary and often plays a vital role in the overall analytical scheme (Lambropoulou et al., 2007). In the past few years, a diversity of extraction methods, including liquid-liquid extraction (LLE) (Fernández-Gutiérrez et al., 1998), stir bar

sorptive extraction (SBSE) (Van et al., 2007), solid-phase micro extraction (SPME) (Beltran et al., 2003) and solid-phase extraction (SPE) (Gil-García et al., 2006), have been developed for sample preparation. A wide range of applications hitherto bears testimony to the fact that they are portable and effective for sample preparation. However, for most of these methods, considerable time and organic solvent are needed for analytes to be extracted into the organic phase or onto the sorbent. In order to increase the throughput for sample pretreatment, the development of time saving and high efficient microextraction techniques has drawn much attention (Ding et al., 2013).

Dispersive liquid-liquid microextraction (DLLME) has been brought up as an exciting microextraction technique in 2006 (Yu et al., 2014; El-Shahawi et al., 2013; Rezaee et al., 2006). In this procedure, a mixture of high density organic solvent (serving as extraction solvent) and water miscible polar dispersive solvent (dispersive solvent) is rapidly injected into an aqueous sample to form an emulsion consisting of fine droplets of the extraction solvent, dispersive solvent, and water. Due to the extraction solvent being highly dispersed in the aqueous phase, the surface area between extraction solvent and sample solution is infinitely large, thus speeding up the extraction. While, the main disadvantage of DLLME is that the extraction solvent is generally limited to solvents with higher density than water in order to be sedimented by centrifugation, typically chlorinated solvents such as chlorobenzene, chloroform, and tetrachloromethane, all of them are potentially toxic to humans and the environment. To date, CPE has already successfully been applied for the quantification of Au-NPs in water samples in combination with electrothermal atomic absorption spectrometry (ET-ASS) (Hartmann et al., 2013). However, ET-AAS measurement, which background absorption signal is large, the relative sensitivity and sample in homogeneity more serious, analytical precision is poor.

In this study, the possibility of AgNPs enrichment by DLLME was considered. A new microextraction method combined with flame atomic absorption spectrometry (FAAS) was developed for separation, enrichment and determination of AgNPs in environmental waters. Factors affecting the extraction efficiency, such as pH, concentration of extractant, extraction time, and nature of the extractant were studied and optimized.

Materials and methods

Materials

All chemicals used were analytical grade or better and all solutions were prepared in deionized water. 1-hexanol, 1-heptanol, 1-octanol, 1-nonanol, 1-decanol and decanoic acid were obtained from the Aladdin Reagent Corporation (Shanghai, China). Tannic acid, sodium carbonate, Ammonium sulfate, hydrochloric acid, silver nitrate, ethanol and sodium thiosulphate, all purchased from (Aladdin, China).

Equipment

A pHSJ-4A pH meter (Shanghai, China), vortex agitator (Shanghai, China) and SK5210LHC ultrasonic instrument (Shanghai, China), centrifuge (Shanghai, China) and flame atomic absorption spectrometer (AA6300, Shimadzu, Japan) were used.

Synthesis of AgNPs nanoparticle

Preparation of $AgNO_3$ solution and tannic acid solution, 0.100 mol/L Na_2CO_3 solution with deionized water; amount of 200 mL tannic acid solution at three mouth bottle

thermostatic water bath heating, and stirring; adding 5 mL Na₂CO₃ solution to tannin reduction; after 10 min, A certain amount of volume AgNO₃ solution, with the constant pressure drop funnel the dropwise added to the above mixture. Three mouth bottle solution color from colorless to light yellow gradually, orange, red, and finally turned a deep red or red brown, continue stirring heating 30 min, obtained the AgNPs. The AgNPs concentration of the solution was determined by FAAS.

Extraction procedure

DLLME of AgNPs step

A 4 mL AgNPs sample solution placed in 10 mL long tapered glass centrifuge tube was adjusted to pH 3.0 with diluted HCl, and then 40 μ L of 1 M Na₂S₂O₃ and 80 μ L of 1-octanol (as extraction solvent of DLLME) were added to the acetonitrile/water (80:20, v/v) extract (used as disperser solvent) and the mixture was diluted to 40 mL by deionized water in a 50 mL vial with conical bottom. Then, the tube was sealed and swirled on a vortex agitator at 2000 rpm for 30 s to accelerate the formation of the fine droplets of the extraction solvent and to enhance the transfer of the analytes. The supernatant aqueous phase can then easily be removed by decanting. The droplet was dissolved in 2 mL of ethanol and transferred to FAAS measurement.

Results and discussion

Optimization of the DLLME conditions

Parameters that commonly affect the DLLME efficiency, such as sample pH, the type and volume of extraction solvent, salt content, equilibration time were investigated.

pH plays an important role on metal chelate formation and subsequent extraction. To optimize the extraction conditions for the nanoparticles pH was varied between 1 and 7 using hydrochloric acid for pH adjustment while the other experimental variables remained constant. The highest extraction efficiency was achieved at pH 3.0 (*Fig. 1*) while the replicate times n = 6. So, pH 3.0 was selected as the optimum for further experiments.



Figure 1. Effect of pH on DLLME of AgNPs

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Selection of the disperser solvent

The applicability of several organic solvents including MeOH, MeCN, EtOH, MeOH 80% and MeCN 80% was investigated in the preliminary experiments taking into account that they should be used as primary extracts of the analytes from real water samples must then act as disperser solvent in DLLME process. *Figure 2* shows the recoveries of the whole procedure of all extraction solvents tested. The results revealed that the maximum extraction efficiency was achieved by MeCN/water (80:20 v/v).



Figure 2. Effect of dispersing solvent

To study the volume effect of disperser solvent, it was varied from 4 to 12 mL in the interval of 2 mL while extraction solvent was kept at 80 μ L. The obtained results showed that the extraction efficiency increased with increasing the volume of MeCN 80% up to 10 mL and then dropped down slightly above 10 mL due to the increased solubility of AgNPs in the aqueous phase and decreasing the distribution ratio.

Therefore 10 mL MeCN 80% was selected as both the extraction solvent of AgNPs from environmental samples and as disperser solvent in DLLME for subsequent experiments.

In this study $Na_2S_2O_3$ as a masking reagent to cover the interference of Ag^+ . When $Na_2S_2O_3$ was added, thiosulfate ions and silver ions generated thiosulfate (*Fig. 3*), which is soluble in water. Therefore, silver ions can be eliminated to the detection of AgNPs.

$$Na_{2}S_{2}O_{3} + 2Ag^{+} = 2Na^{+} + Ag_{2}S_{2}O_{3}\downarrow$$
$$Ag_{2}S_{2}O_{3} + 2Na_{2}S_{2}O_{3} = Ag_{2}Na_{4}(S_{2}O_{3})_{3}$$

Figure 3. The reaction of sodium thiosulfate with Ag^+

Selection of extraction solvent

Selection of appropriate extraction solvent is of great importance in microextraction technique in order to obtain efficient extraction. For the LDS-DLLME, the extraction

solvent should meet the following requirements: (1) be immiscible with aqueous solution (2) able to extract the analytes, and (3) possess low toxicity. To avoid the use of more highly toxic solvents such as trichloromethane, tetrachloromethane, tetrachloromethylene, chlorobenzene etc., five low density organic alcohol including 1-hexanol (density, $d = 0.819 \text{ g mL}^{-1}$), 1-heptanol ($d = 0.820 \text{ g mL}^{-1}$), 1-octanol ($d = 0.821 \text{ g mL}^{-1}$), 1-nonanol ($d = 0.827 \text{ g mL}^{-1}$), and 1-decanol ($d = 0.828 \text{ g mL}^{-1}$) were evaluated as extractants. Among them a stable cloudy solution and good extraction recovery were observed with 1-octanol. The fine droplets of 1-octanol afford larger surface area to contact with the water sample and can be completely absorbed by MNPs after magnetic separation, so that higher extraction performance can be attained. As can be seen, 1-octanol has higher extraction recovery than those obtained by other extraction solvents for the AgNPs.

Effect of extraction solvent volume

The volume of extracting solvent is an important parameter which can influence the occurrence of the cloudy state and efficiency of extraction process. To investigate the volume effect of 1-octanol, different quantities of 1-octanol (50-100 μ L) were used. Observably, too little 1-octanol may lead to low recoveries and the extraction efficiency was improved with the increase of 1-octanol volume. As shown in *Figure 4*, the recoveries for each of the AgNPs increased from 55-63% to 86.8-109.1% in the range of 50-80 μ L of 1-octanol. The highest recovery was obtained when 80 μ L of 1-octanol was selected as an optimum solvent volume for further studies.



Figure 4. Selection of extraction solvent

Effect of equilibration time

In DLLME extraction, equilibration time is defined as an interval time from the occurrence of cloudy solution and before addition of magnetic adsorbents. The effect of extraction time on the adsorption was investigated from 0-120 s. The result showed that

equilibration time has no influence on extraction efficiency. It is well-known that vortex process is an effective way to enhance mass transfer from aqueous phase to the extraction phase. Consequently, short time was required for extraction. Generally, 30 s of extraction time was enough to achieve high extraction recovery, and prolonged extraction time did not contribute significantly to an increase in extraction recovery. Thereby, vortex time of 30 s was selected.

The high viscosity of the so prepared surfactant rich phase made it necessary to dissolve the sample prior to sample introduction to FAAS. Ethanol appeared to be a feasible solvent since it readily dissolves the surfactant rich phase and it prevents degradation of the enriched nanoparticles.

Evaluation of the preservation of the nanoparticle size distribution

To gain an insight in the size distribution of the nanoparticles in real samples it is necessary to know whether the size distribution of Ag-NPs is changing during the extraction procedure or not. This was studied using TEM for imaging of the particles before and after DLLME. As shown in *Figure 5*, the size of particles that underwent the extraction process is more or less identical to the size before DLLME treatment. Thereby we conclude that the size distribution is not affected by the presented method. This supports the results of Yang et al. (2017) that studied CPE for metal nanoparticles. This finding is especially important because the size determination of nanoparticles is an important issue in current research. DLLME samples of Ag-NPs can therefore be used for size determination after pre-concentration.



Figure 5. TEM images of synthesized AgNPs

Analytical features

The analytical characteristics of the method were evaluated under the optimum experimental conditions (*Table 1*). The enrichment factor (EF), relative standard deviations (RSD) and limits of detection (LOD) are effective factors to evaluate a developed analytical method. With the optimized system, the calibration graph of AgNPs was linear in the range of 10-1000 μ g/L. The calibration equation is A = 7.9 × 10⁻² C + 0.003 with a correlation coefficient of 0.9999, where A is the absorbance and C is the Ag concentration in μ g/L. The detection limits, defined as the

concentration equivalent to three times the standard deviation of the reagent blank, for AgNPs was 6.5 μ g/L. The relative standard deviation for AgNPs was 2.1% (n = 9). The relative recoveries of AgNPs in water samples at the spiking level of 3-20 μ g/L ranged from 96.7 to 108%. There are many reasons, such as instrument fluctuations, operating errors, etc., so the calibration curve will appear quantitative. Recovery rates of 80% to 120% are generally considered normal. The enhancement factor, defined as the ratio of slope of preconcentrated samples to that obtained without preconcentration, was 57 for AgNPs.

Parameter	Analytical feature		
Linear range (µg/L)	10-1000		
\mathbf{R}^2	0.9999		
Limit of detection (µg/L)	6.5		
RSD (%) (n = 9)	2.1		
Enrichment factor	57		
Pre-concentration time (min)	<10		
Recovery (%)	96.7-108.0		

Table 1. Analytical characteristics of the method

Effect of foreign ions

To regard of the high sensitivity of FAAS, the interference connected with the DLLME procedure was studied. Due to cations may react with AgNPs and lead to the decrease of extraction efficiency.10 mL sample solution containing Ag^+ and other ions were prepared and treated with the developed procedure. The tolerance limit was defined as the foreign ion concentration causing a change in the absorbance of less than $\pm 5\%$. *Table 2* shows the concentration of the interferential ions, the recovery is tolerable.

Table 2. Effects of the matrix ions on the recoveries of the examined metal ions

Ions		Recovery (%)		
	Concentration (ug/L)	Pd	Ag	
Na ⁺	10×10^{3}	105.0	104.1	
\mathbf{K}^+	10×10^{3}	104.2	104.8	
Cl^-	10×10^{3}	103.5	102.2	
Ca ²⁺	10	99.1	101.1	
Mg^{2+}	0.5	99.2	96.8	
Cr^{3+}	0.5	100.7	98.8	
Mn^{2+}	0.5	103.4	99.2	
Zn^{2+}	0.5	102.2	100.1	
Cd^{2+}	0.5	98.7	99.7	
Ni ²⁺	0.5	99.6	100.8	
Pb^{2+}	0.5	108.7	98.5	
Cu^{2+}	0.5	102.6	99.9	

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Analysis of real samples

To further evaluate the applicability of the proposed method, real environmental water samples were analyzed and the recoveries were determined by spiking 3-20 μ g/L AgNPs. The WWTP water sample was obtained from our laboratory (Bijie, China), the river water sample from LiuCang Qiao River (Bijie, China). As shown in *Table 3*, the obtained recoveries of AgNPs were in the range of 96.7-108%, which was satisfactory with regard to the low spiking level. No significant difference between recoveries for samples with or without filtration was observed, suggesting that the presence of natural particles in the samples had no significant effects on the DLLME of AgNPs at concentrations 20 μ g/L. This result indicates that FAAS can be adopted to identify and quantify the AgNPs content in environmental samples.

Table 3. Determination and identification of AgNPs in environmental waters by the proposed method

Sample	AgNPs added (µg/L)	AgNPs found (µg/L)	Recovery (%)
River water	0	*ND	_
	5	4.9±0.3	$98.0{\pm}0.6$
	10	10.8±0.5	108.0 ± 5.0
	20	20.5±0.6	102.5±3.0
WWTP	0	*ND	_
	3	2.9±0.5	96.7±16.7
	8	8.1±0.6	101.3±7.5
	15	15.8±0.7	105.3±4.6

Comparison with other measurements

Table 4 compares the characteristic data of the proposed method with other methods for AgNPs reported in literature. As it can be seen, the proposed method has higher RSD and lower detection time.

Measurement	Sample volume (mL)	Detection time (min)	RSD (%) (n = 9)	Injection volume (mL)	LOD (µg/L)	Reference
ICP-MS	9.5	>10	5.6	5	0.006	Hartmann et al. (2013)
ET-AAS	40	5	9.5	0.2	0.005	Hartmann et al. (2013)
FAAS	5	<2	2.1	0.2	6.5	This paper

Table 4. The comparison of proposed measurement with other measurements

Conclusion

In the present study, a new microextraction technique, based on dispersive liquidliquid microextraction (DLLME) was developed for determining AgNPs in environmental water samples. $Na_2S_2O_3$ was added into the extraction system as a masking reagent of Ag⁺ to prevent its extraction into the 1-octanol rich phase. The extraction could be achieved within 2 min and good extraction efficiencies were obtained. Such facile separation is essential to improve the operation efficiency which avoiding time-consuming centrifugation or freezing or manual collection of extractant. In addition, there are many reasons, such as instrument fluctuations, operating errors, etc., so the calibration curve will appear quantitative. Recovery rates of 80% to 120% are generally considered normal. The type of sample and matrix does not affect the detection of AgNPs. This proposed method provided a simple approach for quantification of AgNPs in environmental water samples.

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