SPATIOTEMPORAL ASSESSMENT OF NUTRIENTS IN THE STEELPOORT SUB-CATCHMENT OF THE OLIFANTS RIVER BASIN, SOUTH AFRICA

ADDO-BEDIAKO, A.

Department of Biodiversity, University of Limpopo, Private Bag X1106, Sovenga, 0727, South Africa e-mail: abe.addo-bediako@ul.ac.za; phone: +27-15-268-3145; ORCID: 0000-0002-5055-8315

(Received 22nd Jun 2023; accepted 22nd Aug 2023)

Abstract. Nutrient enrichment is considered to be one of the most serious threats to freshwater ecosystems. The increasing agricultural and other human activities in the Steelpoort sub-catchment of the Olifants River Basin has led to serious impacts on water quality. The spatial and seasonal variability of nutrient concentrations in the Dwars, Spekboom and Steelpoort rivers in South Africa were analyzed. Four monitoring sites per each river were selected to assess the impact of human activities on nutrient levels in the rivers during the wet and dry seasons. The samples were analyzed to determine the concentrations of nitrates, nitrites, ammonia and ortho-phosphates. Trends in nutrient concentrations varied among the sites and rivers, and there were significantly higher levels of nutrients detected at sites with high human activities. Generally, higher concentrations of nutrients were recorded during the wet season in the Dwars and Steelpoort rivers and higher concentrations of nutrients were recorded in the Spekboom River during the dry season. The most likely cause of increasing nutrient enrichment are effluent from agricultural fields, partially treated and untreated sewage. Though nutrient levels recorded were low, some sites had concentrations that exceeded the recommended limits. It is therefore important that proper management measures should be implemented to reduce wastes discharged into the rivers. **Keywords:** *land use, eutrophication, nutrients, pollution, water quality*

Introduction

Globally, water quality is declining due to increasing urbanization, mining, agricultural and industrial activities (Cheng et al., 2021; Addo-Bediako and Rasifudi, 2021; Varol et al., 2022). These activities, coupled with channel alteration and climate change are causing a decline in water quality in freshwater ecosystems, and are affecting the biotic communities (Liu et al., 2021; Ge et al., 2022). Nutrient enrichment is one of the major problems freshwater ecosystems are facing, usually caused by human activities (Romanelli et al., 2020). A number of studies have expressed concern about the impact of nutrient enrichment (eutrophication) on freshwater resources (e.g. Van Ginkel, 2011; Harding, 2015; Gqomfa et al., 2022). Nutrients, especially nitrogen and phosphorous generally originate from over fertilized agricultural soils through leaching and erosion processes (Trinh et al., 2015; Singkran, 2017) or from untreated domestic and industrial wastewaters (Cloern et al., 2014).

Nutrients can be stored temporarily in the groundwater and then transported into the stream via subsurface water or surface runoff that is induced by rainfall (Laurent and Mazumder, 2014; Zhou et al., 2017). High nutrient content can lead to eutrophication, may cause depletion of dissolved oxygen, and lead to increased autotroph biomass, species compositional shifts, reductions in biodiversity, potential production of algal toxins, and affect the taste and odor of water (Griffin, 2017). Assessment of the ecological status of rivers is important in managing river ecosystems and providing a precondition for economic growth and ecological integrity (Cao et al., 2019).

In South Africa, nutrient enrichment in freshwater ecosystems is increasing due to increasing runoff from agricultural activities, industrial wastewater, and municipal sewage due to dysfunctional water waste treatment plants (Dabrowski and De Klerk, 2013). Due to the increasing nutrient levels in many rivers, there are concerns about severe ecological and economic consequences (Harding, 2015). The Olifants River Basin is one of the most polluted river systems in South Africa and has been systemically impaired by increasing agricultural, industrial and urban development, resulting in contamination of the water (Ashton and Dabrowski, 2011). Intensive and subsistence agriculture activities in the Steelpoort sub-catchment of the Olifants River Basin, in conjunction with industrial and other human activities are negatively impacting on the water quality of the rivers (Ashton and Dabrowski, 2011; Addo-Bediako, 2021). Though, there are a number of programs to monitor nutrient levels, there is relatively little published information on nutrient levels of freshwater ecosystems in the Steelpoort sub-catchment of the Olifants River Basin. Thus, the aim of this study was to assess the spatial and seasonal patterns in nutrient levels and to evaluate how nutrient levels are related to the land use in the river catchments.

Materials and methods

Study area

The Steelpoort River sub-catchment of the Olifants River Basin is located at the north eastern part of South Africa. The main rivers in the sub-catchment include Steelpoort, Dwars and Spekboom rivers. The rivers provide water for agricultural, mining and industrial activities. They also serve as a source of drinking water to a number of communities in the area. The Steelpoort River sub-catchment covers an area of 7,139 km². The elevation ranges from 1,500 m to 2,400 m, with the exception of the Steelpoort River valley, which is between 900 and 1,200 m. The mean annual rainfall for the area is between 630-1,000 mm. Rainfall usually occurs in the summer months, from October to March, with the highest rainfall usually in January. The mean daytime summer temperatures are between 19°C and 22°C, while the winter mean temperatures are between 13°C and 19°C (Stimie et al., 2001; SAWS, 2023).

Stream water sampling and chemical analysis

Twelve sampling sites were selected along the three main rivers in the Steelpoort River sub-catchment, the Dwars, Spekboom and Steelpoort rivers (*Table 1; Fig. 1*). Seasonal sampling was conducted during wet and dry seasons of 2019 and 2020. Stream water was collected in polyethene bottles from each of the 12 sites. Water samples collected from the streams were placed on ice and transported to the laboratory. They were filtered within 24 h of collection using 0.7-mm glass microfiber filters. The water samples were then preserved in polyethene bottles at 4°C prior to chemical analysis. The nitrate $(NO₃)$, nitrite $(NO₂)$, ammonia $(NH₃)$ and ortho-phosphate $(PO₄)$ levels were determined using a spectrophotometer (Merk Pharo 100 Spectroquant™) with Merck cell test kits. Physicochemical parameters such as water temperature, dissolved oxygen (DO), pH, electrical conductivity (EC) and total dissolved solids (TDS) were measured in situ using YSI multi-parameter instrument (Yellow Springs Instruments, Texas, model Model 554 Data logger).

Sites	Latitude	Longitude	Description of the sites		
DWS1	-24.8553	30.0103	Upstream, near mine water return dam and grazing area		
DWS ₂	-24.8428	30.0867	Midstream, near mining and industrial areas (ferrochrome mine and smelter)		
DWS3	-24.8317	30.0797	Midstream, area with agricultural coverage and pasture		
DWS4	-24.8306	30.0794	Downstream, near the mouth, agricultural activities (crop production and animal grazing) and informal settlements		
SPS ₁	-24.6936	30.3628	Upstream reach, agricultural and mining areas		
SPS ₂	-24.6581	30.3386	Midstream, water treatment plant		
SPS3	-24.6586	30.3244	Midstream, wastewater treatment plant		
SPS4	-24.6419	30.3081	Downstream, near human settlement and grazing area		
STS ₁	-25.1058	29.8519	Upstream, near grazing area		
STS ₂	-24.9942	30.0186	Midstream, near recreational site, site used for washing of clothes, cars and fishing		
STS3	-24.8836	30.1192	Midstream, new residential development		
STS4	-24.7339	30.5011	Downstream, near grazing area		

Table 1. Characterization of selected sampling sites in the Steelpoort sub-catchment of the Olifants River Basin

DWS = Dwars River Sites, SPS = Spekboom River S STS = Steelpoort River Sites

Figure 1. Map of the study area, showing the sites of the Dwars, Spekboom and Steelpoort rivers (DWS = Dwars River Sites, SPS = Spekboom River S STS = Steelpoort River Sites)

Statistical analysis

The mean and standard deviation of the physicochemical parameters and nutrients were determined. One-way analysis of variance (ANOVA) was used to determine differences for mean concentrations of nutrients and dissolved oxygen among sites and seasons. The occurrence of a linear correlation between analyzed variables was tested using non-parametric Spearman Rank Correlation Analysis. The correlation determines the degree of a linear relationship between the different indicators. The level of significance considered for the interaction was 0.05. All statistical analyses were done using Statistica (Version 10).

Results

Physicochemical parameters

The water quality parameters showed variations among the rivers and along the rivers as shown in *Table 2*. Surface water temperature ranged from 17.6 to 22.2°C in the Dwars River, from 18.78 to 23.63°C in the Spekboom River, and from 17.4 to 25.75°C in the Steelpoort River. The lowest mean temperature was 18.8 in the Steelpoort River and the highest mean temperature was 21.7 in the Spekboom River (*Table 3*). The pH of the Dwars River ranged from 7.55 to 8.76, the pH of the Spekboom River ranged from 6.68 to 8.25, and the pH of the Steelpoort ranged from 7.80 to 8.85. The highest pH values were recorded during the wet season in all the rivers. The electrical conductivity values ranged from 282.4 to 559.6 uS/cm in the Dwars River, from 282.4 to 628.8 uS/cm in the Spekboom River, and from 232.9 to 372.5 uS/cm in the Steelpoort River. The TDS levels were between 96.5 and 385.5 mg/l in the Dwars River, between 193.6 and 436.5 mg/l in the Spekboom River, and from 181.6 to 240.5 mg/l. Generally, EC and TDS levels were higher during the wet season than the dry season in all the three rivers, with the highest mean EC and TDS of 515.8 uS/cm and 335.5 m/l respectively (*Table 3*).

The DO concentration in the Dwars River ranged from 5.32 to 8.65 mg/l, from 5.77 to 8.99 mg/l in the Spekboom River, and between 7.52 and 7.84 mg/l in the Steelpoort River. Dissolved oxygen levels fluctuated among the rivers and seasons, with the highest concentration of 8.99 mg/l at S2 in the Spekboom River during the dry season and lowest concentration of 5.32 mg/l at S3 of the Dwars River also during the dry season. The highest mean DO was 8.33 mg/l and the lowest was 6.15 mg/l (*Table 3*). Generally, the DO levels were higher during the dry winter season than in the wet summer season in all the three rivers (*Table 2*).

Spatiotemporal nutrient concentrations

The nutrient concentrations at the sites are shown in *Table 2* and the mean concentrations of nutrients recorded during the wet and dry seasons in each river are shown in *Table 3*. The levels of the nutrients fluctuated greatly among the rivers and between the seasons. Nitrate was recorded in all the sites during both the wet and dry seasons in the three rivers. Nitrate concentration showed an increase during the wet season in the Dwars and Steelpoort rivers, but showed an increase during the dry season in the Spekboom River. The highest mean concentration of 18.5 mg/l was recorded at DWS1 of the Dwars River during the wet season and the lowest concentration of 0.25 mg/l was recorded at STS1 in the Steelpoort River during the dry season. In the Dwars

River, NO² was recorded at all sites during the dry season but only recorded at DWS1 during the wet season. In the Spekboom River, $NO₂$ was below detection level at all sites during the wet season but was detected at all sites during the dry season, with the highest mean concentration of 0.24 mg/l at SPS4. Whilst in the Steelpoort River, NO₂ was below detection level at all sites during the dry season, but was recorded at STS2 during the wet season with a mean of 0.065 mg/l. The highest concentrations of $NO₂$ were recorded during the dry season in the Spekboom River and during the wet season in the Dwars and Spekboom rivers.

Ammonia concentration was below detection level at all sites in the Dwars River during the wet season but was detected at all sites during the dry season with a mean concentration of 0.04 mg/l. In the Spekboom River, NH₃ concentration was below detection level at all sites during the wet season but was detected at all sites during the dry season, with the highest mean concentration of 0.09 mg/l. In the Steelpoort River, however, NH₃ was recorded at all sites during both wet and dry seasons except at S4 during the wet season. In general, the dry season had the highest levels of $NH₃$ in the Dwars and Spekboom rivers, but during the wet season, the highest level of NH³ was in the Steelpoort River. Ortho-phosphate was only detected in the Dwars and Spekboom rivers at all sites during the dry season and was not detected in the Steelpoort River during both seasons.

Site	Temp	$DO(mg/\ell)$	pH	EC	TDS	NO ₃	NO ₂	NH ₃	$Ortho-PO4$
$DWS1-W$	22.2	8.65	7.74	559.6	379	18.5	0.42	n.d	n.d
DWS2-W	17.65	6.68	7.93	514.5	385.5	17	n.d	n.d	n.d
DWS3-W	22.05	6.54	8.14	557.5	382.6	15.5	n.d	n.d	n.d
DWS4-W	21.75	7.63	7.55	282.4	194.7	0.55	n.d	n.d	n.d
DWS1-D	20.4	6.99	8.76	519.5	175.4	12.5	0.25	0.04	0.04
DWS2-D	17.6	5.95	8.2	491.3	181.9	10.4	0.03	0.04	0.05
DWS3-D	20.65	5.32	8.52	542	188.4	12.2	0.03	0.04	0.04
DWS4-D	19.8	6.33	8.35	307	96.5	4.6	0.02	0.04	0.04
SPS1-W	23.63	8.6	6.68	269.4	193.6	0.58	n.d	n.d	n.d
SPS2-W	21.8	8.53	8.16	414.1	277.3	0.3	n.d	n.d	n.d
SPS3-W	20.65	5.77	7.69	628.8	436.5	1.0	n.d	n.d	n.d
SPS4-W	20.65	7.49	7.92	305.2	215.2	0.9	n.d	n.d	n.d
SPS1-D	18.78	8.9	8.18	511.3	309.1	0.4	0.02	0.09	0.06
SPS2-D	21.5	8.99	8.12	485.1	251.2	0.9	0.03	0.03	0.09
SPS3-D	19.7	6.9	8.12	522.2	276.2	1.5	0.04	0.09	0.06
SPS4-D	20.75	8.53	8.25	544.6	272.6	3.9	0.24	0.2	0.11
STS1-W	25.75	7.515	8.65	232.9	198	0.45	n.d	0.45	n.d
STS2-W	21.4	6.9	8.85	293	211.3	0.5	0.065	0.4	n.d
STS3-W	20.25	6.95	8.6	331.5	240.5	1.3	n.d	0.45	n.d
STS4-W	19.1	7.41	7.9	309.2	237.9	1.05	n.d	n.d	n.d
STS1-D	18.45	7.825	8.65	372.5	226.8	0.25	n.d	0.2	n.d
STS2-D	19.8	7.8	8.0	267.6	181.6	0.3	n.d	0.1	n.d
STS3-D	19.35	7.84	7.8	291.7	208.3	0.55	n.d	0.15	n.d
STS4-D	17.4	7.5	8.45	325.4	239.1	0.8	n.d	0.1	n.d

Table 2. Physicochemical parameters taken at each site of the three rivers

DWS-D = Dwars River Sites during dry season, DWS-W = Dwars River Sites during wet season, SPS-D = Spekboom River Sites during dry season, SPS-W = Spekboom River Sites during wet season, STS-D = Steelpoort River Sites during dry season, STS- $W = Steelpoort River Sites during dry season$

> APPLIED ECOLOGY AND ENVIRONMENTAL RESEARCH 22(3):2411-2423. http://www.aloki.hu ● ISSN 1589 1623 (Print) ● ISSN 1785 0037 (Online) DOI: http://dx.doi.org/10.15666/aeer/2203_24112423 © 2024, ALÖKI Kft., Budapest, Hungary

Site	Temp $(^{\circ}C)$	$DO(mg/\ell)$	pН	EC μ S/cm	TDS (mg/l)		$NO3(mq/l)$ $NO2(mq/l)$	NH ₃ (mq/l)	$Ortho-PO4$ (mg/l)
DWS-W		20.9 ± 2.2 7.37 ± 0.98	7.55-8.14	478.5 ± 132	335.5 ± 94		$12.90 \pm 8.3 \pm 0.11 \pm 0.21$	0.0	0.0
DWS-D	19.6 ± 1.4	6.15 ± 0.7	8.20-8.76	465.0 ± 107	160.6 ± 43	9.92 ± 3.7	0.08 ± 0.1	0.04 ± 0.0	0.04 ± 0.01
SPS-W		21.7 ± 1.4 7.60 \pm 1.32	$6.70 - 8.20$	404.4 ± 162	280.7 ± 110	0.7 ± 0.31	$0.0\,$	0.0	0.0
SPS-D	20.2 ± 1.2	8.33 ± 0.97	8.11-8.25	515.8 ± 25	277.3 ± 24	1.68 ± 1.6	0.08 ± 0.11	0.10 ± 0.07	0.08 ± 0.02
STS-W	21.6 ± 2.9	7.19 ± 0.3	7.9-8.90	291.7 ± 42	221.9 ± 21	0.83 ± 0.42	0.2 ± 0.03	0.33 ± 0.2	0.0
STS-D	18.8 ± 1.1	7.74 ± 016	$7.8 - 8.65$	314.3 ± 45	214.0 ± 25	0.48 ± 0.25	0.0	0.14 ± 0.05	0.0
CCME			$6.0 - 9.0$			13.0	0.06	0.354	
SANS			$6.5 - 9.5$			11.0	< 0.9	< 1.5	
WHO	< 25.0		$6.5 - 8.5$	600	400.0	50.0	3.0	0.2	0.1

Table 3. The mean ± standard deviation (SD) of physicochemical parameters measured in the Dwars, Spekboom and Steelpoort rivers of the Olifants River Basin

DWS-D = Dwars River Sites during dry season, DWS-W = Dwars River Sites during wet season, SPS-D = Spekboom River Sites during dry season, SPS-W = Spekboom River Sites during wet season, STS-D = Steelpoort River Sites during dry season, STS-W = Steelpoort River Sites during dry season

Differences in the nutrient and dissolved oxygen concentrations among the rivers

The $NO₃$ and $NH₃$ concentrations among the rivers differed significantly during the wet season (F = 8.481, $p < 0.005$; F = 8.895, $p < 0.005$ respectively) (*Fig. 2*), and during the dry season, NO3, Ortho-PO⁴ and DO differed significantly among the three rivers $(F = 105.87, p < 0.0001; F = 30.76, p < 0.0001; F = 10.45, p < 0.005$ respectively) (*Table* 4). The post-hoc tests showed that the $NO₃$ difference in mean concentration was between Dwars and Spekboom rivers ($p = 0.0002$) and between Dwars and Steelpoort rivers ($p = 0.0001$), the difference in NH₃ concentration was between Dwars and Steelpoort rivers ($p = 0.003$), between Spekboom and Steelpoort rivers ($p = 0.011$), the difference in ortho-PO⁴ concentration was between Spekboom and Steelpoort rivers $(p = 0.033)$, and the difference in DO concentration was between Dwars and Spekboom rivers ($p = 0.039$).

Nutrients	df		F	p-value			
Wet season							
NO ₃	$\overline{2}$	196.14	8.481	0.008			
NO ₂	$\overline{2}$	0.013	0.849	0.459			
NH ₃	$\overline{2}$	0.141	8.895	0.007			
Ortho-PO ₄	$\overline{2}$						
DO.	$\overline{2}$	0.162	0.174	0.843			
Dry season							
NO ₃	$\overline{2}$	105.87	19.94	0.0005			
NO ₂	$\overline{2}$	0.009	1.154	0.358			
NH ₃	$\overline{2}$	0.010	4.001	0.057			
Ortho-PO ₄	$\overline{2}$	0.006	30.76	< 0.0001			
DO	2	5.089	10.45	0.005			

Table 4. Results of analysis of variance among the rivers during wet and dry seasons

Figure 2. Box and whisker plots for the nutrient and dissolved oxygen distribution in the Dwars, Spekboom and Steelpoort rivers (DWR = Dwars River, SPR = Spekboom River, STR = Steelpoort River)

In the Dwars River, the highest concentrations of $NO₃$ was recorded at DWS1 (upstream) and the lowest at DWS3 (midstream), the highest mean $NO₂$ was at DWS1 and the lowest at DWS4 (downstream), however, the concentrations of NH₃ and ortho-PO₄ were similar at all reaches of the river, and the highest DO was at DWS1 and the lowest mean concentration was at DWS3. In the Spekboom River, NO3, NO2, NH3, ortho-PO4 concentrations were highest at SPS4 (downstream), and the highest mean concentration of DO was at SPS2. In the Steelpoort River, the highest mean NO₃ was at STS3 and STS4, NO² was only detected at STS2. The concentrations of NH³ and ortho-PO⁴ were similar at all sites of the river, and the highest DO was at STS1, the highest mean concentration of NH³ was at STS1. Ortho-PO4 was below detection level at all sites and the highest mean DO was at STS1. There was a strong positive correlation between $NO₃$, and ortho- $PO₄$ (0.705) , a significant correlation between $NO₃$ and $NO₂$, and a significant negative correlation between NO³ and DO, thus, DO decreases with increasing NO³ (*Table 5*).

Table 5. Correlation matrix of water quality indicators in the wet and dry seasons in the Steelpoort sub-catchment of the Olifants River Basin

	NO ₃	NO ₂	NH ₃	PO ₄	D _O
NO ₃	1.000000				
NO ₂	0.442885	1.000000			
NH ₃	-0.310360	0.043718	1.000000		
PO ₄	0.279588	0.704535	0.073974	1.000000	
D _O	-0.503351	0.005593	0.026450	0.024657	1.000000

Discussion

Physicochemical parameters

Water quality has a significant effect on the aquatic biota (Ali et al., 2016). The mean water temperature was found within the permissible limits of 25°C (WHO, 2011) at all the sites of the three rivers. The pH was slightly alkaline at all sites of the three rivers. The high values of electrical conductivity and TDS recorded in the present study were expected owing to the level of human activities in the catchments of the rivers. Conductivity and TDS attained their highest levels in the midstream (SPW3) of the Spekboom River during the wet season. An increase in conductivity and TDS in the Spekboom River can be attributed to anthropogenic discharges and runoff of wastewater into the river. The EC and TDS levels are usually increased by anthropogenic factors such as, land use changes, domestic, industrial, agricultural (intensive fertilization, irrigation) discharges (Canedo-Arguelles et al., 2013; Kaushal et al., 2018). High waste discharges into water bodies can make the water less suitable for its intended purposes (Islam and Islam, 2020), unfortunately, all these activities occur along the three rivers in the Steelpoort sub-catchment of the Olifants River Basin.

Dissolved oxygen concentration in water bodies is important in supporting aquatic life. In general, the DO was highest in the Spekboom River, followed by the Steelpoort River and then the Dwars River. From the results, with the decrease in water temperature during the dry winter season, the concentration of DO increased. Thus, the DO concentration reached its maximum value in the Spekboom River (8.99 mg/l) and the Steelpoort River (7.84 mg/l) during dry winter seasons. The high DO in dry winter season besides the decrease in temperature could be due to a decrease in the rate of oxygen consumption from aquatic organisms (Ali et al., 2016). However, in the Dwarf River, the highest DO (8.68 mg/l) was recorded during the summer wet season and the lowest DO concentration (5.32 mg/l) was recorded during the dry season. The low DO concentration in the Dwars River during dry winter season could be attributed to the high nutrient content ($NO₃$ and $NO₂$) and the slow-moving water.

Spatiotemporal stream nutrient concentrations

The higher $NO₃$ concentration in the Dwars and Steelpoort rivers during the wet season than dry season could be attributed to runoff from agricultural fields in the

upstream of the river. High levels of $NO₃$ are normally due to improper disposal of solid waste, sewage and extensive use of chemical fertilizers (Garzon-Vidueira et al., 2020). The results of the study are consistent with other studies in the region which found higher $NO₃$ concentrations during the wet than dry seasons (Nyamangara et al., 2013; Edokpayi et al., 2015). Intense rainfall events usually lead to erosion and surface runoff transporting suspended materials into the water bodies that can increase nutrient load in rivers (Deng et al., 2021). The level of nutrients is usually high in areas such as recently fertilized fields and shallow nutrient enriched groundwater (Jiang et al., 2014; Orem et al., 2015). High $NO₃$ concentration during wet season can also be caused by nitrification process which causes a rapid decrease of $NH₃$ and an increase of $NO₃$ (Nguyen et al., 2019). Thus, there is seasonal variations in the transport of nitrates into water bodies.

In the Spekboom River, however, the $NO₃$ concentration during the dry season was higher than during the wet season. This could be due to partially treated sewage discharged into the river from the dysfunctional waste water treatment plant at site SPS3, which also affected the downstream of the river (SPS4). Wastewater discharge from rural and agricultural areas constitutes a constant source of pollution in many freshwater bodies. Despite the increasing nutrient pollution in some parts of the rivers, the NO3 levels during both the dry and wet seasons were within the standard guideline values (CCME, 2012; WHO, 2011; SANS, 2015), however, the mean concentration of NO³ exceeded SANS (2015) value in the Dwars River. Dorleku et al. (2019) reported that $NO₃$ at concentrations between 10 to15 mg/l can affect the water quality, aquatic biota and even humans. In infants, high $NO₃$ levels in drinking water can cause methemoglobinemia, a condition where the blood is unable to bind and carry oxygen to the cells (Nartey et al., 2012).

The $NO₂$ concentrations were consistently below the lower detection limit $(<0.01$ mg/L) at many sites. At sites where $NO₂$ was detected, the levels were below the guideline values of SANS (2015) and WHO (2011) of 0.9 and 3.0 mg/l respectively. The absence or low levels of $NO₂$ in the rivers could be attributed to the fact that $NO₂$ is a pervasive intermediate in the nitrogen cycle in nature. The highest $NH₃$ concentration was recorded in the Steelpoort River during the wet season. The levels of NH₃ exceeded the WHO standard guideline value of 0.2 mg/l, at the upstream and midstream sites (STS1, STS2, and STS3) of the Steelpoort River during the wet season and the NH³ concentration exceeded the South African National Standard guideline of < 1.5 mg/l at SPS3 of the Spekboom River (SANS, 2015). The increase mainly during the wet season may be attributed to waste from farming activities and sewage entering the river through runoffs. The presence of high concentration of NH³ usually indicates recent contamination by decaying organic matter, while the presence of $NO₃$ is indicative of late pollution (Gradilla-Hernández et al., 2020). High levels of NH₃ may trigger environmental problems, such as nitrous oxide emission and depleted oxygen in the water.

Ortho-phosphate is an essential nutrient for the development of organisms, and it is not categorized as a harmful or toxic element at low concentrations (Kumar and Puri, 2012). The natural source of phosphorus compounds in freshwater ecosystems is due to the dissolution of soil compounds, the decomposition of organic matter and the breakdown of microorganisms. Anthropogenic sources are related to domestic and industrial wastes, detergents, animal excrement and fertilizers (Griffin, 2017; Zhou et al., 2017). Ortho-phosphate was only recorded in the Dwars and Spekboom rivers during the dry season. There was not much fluctuation of ortho- $PO₄$ levels in the two rivers, with the highest recorded at SPS4 (downstream) of Spekboom River. During the dry season, the Spekboom River receives water discharge from the WWTP at SPS3 and could affect water downstream. Discharges from municipal and industrial wastewater treatment plants (WWTPs) are primary point sources of nutrients to receiving freshwater systems. Despite inputs being less diluted during the dry season, the decline in ortho-PO⁴ concentration in the Dwars and Steelpoort rivers could also be attributed to algal growth (Causse et al., 2015; Nguyen et al., 2019).

The results are consistent with other studies which observed that land cover changes such as agricultural activities are a major driver of changes in water quality, mainly during the wet season (Mello et al., 2018; Alves et al., 2019). The enrichment of waters with ortho-phosphate and nitrogen compounds is an indication of runoff of enriched soils into the water bodies (Camara et al., 2019). The nutrients recorded in the rivers are mainly from agricultural runoff (wastes, pesticides and fertilizers), domestic (sewage and detergents) and industrial wastewater (Fernandes et al., 2021; Olive, 2021).

Conclusion

Pollution originating from human activities had a strong impact on nutrient levels $(NO₂, NO₃, NH₃, ortho-PO₄)$ and DO in the three rivers. Concentrations of $NO₂, NO₃$, NH₃ and ortho-PO₄ in the Spekboom River were lower during the wet season, likely due to greater dilution from the higher rainfall. In contrast, $NO₂$, $NO₃$ concentrations were higher during the wet season in the Dwars River and $NO₂$, $NO₃$, $NH₃$, were higher in the wet season in the Steelpoort River. Thus, there is a clear separation among the dry and wet seasons, showing the strong influence of the seasonality on the variations of nutrients in the sub-catchment. The results suggest that large amounts of agricultural runoff reach the Dwars and Steelpoort rivers during the wet season, while in the Spekboom River, during the dry season, excess nutrients are received from the dysfunctional wastewater treatment plant in the midstream of the river. Excess nutrient in the rivers can reduce habitat availability, increase nuisance plant/algae growth, and decrease dissolved oxygen levels in the water. The implementation of good land use practices in the catchment is necessary to reduce erosion and nutrient discharge into the rivers. Furthermore, frequent monitoring to check the levels of nutrient pollution is needed for a swift action.

REFERENCES

- [1] Addo-Bediako, A., Rasifudi, L. (2021): Spatial distribution of heavy metals in the Ga-Selati River of the Olifants River System, South Africa. – Chemistry and Ecology 37(5): 450-463.
- [2] Ali, M. M., Ali, M. L., Islam, M. S., Rahman, M. Z. (2016): Preliminary assessment of heavy metals in water and sediment of Karnaphuli River, Bangladesh. – Environmental Nanotechnology, Monitoring & Management 5: 27-35.
- [3] Alves, W. S., Martins, A. P., Morais, W. A., Pôssa, E. M., Castro, R. M., Moura, D. M. B. (2022): USLE modelling of soil loss in a Brazilian Cerrado catchment. – Remote Sensing Applications: Society and Environment 27: 100788.
- [4] Ashton, P. J., Dabrowski, J. M. (2011): An Overview of Water Quality and the Causes of Poor Water Quality in the Olifants River Catchment. WRC Project No. K8/887. – Water Research Commission, Pretoria.
- [5] Camara, M., Jamil, N. R., Abdullah, A. F. B., Hashim, R. B. (2019): Spatiotemporal assessment of water quality monitoring network in a tropical river. – Environmental Monitoring and Assessment 191: 729.
- [6] Canedo-Arguelles, M., Kefford, B. J., Piscart, C., Prat, N., Schafer, R. B., Schulz, C-J. (2013): Salinisation of rivers: an urgent ecological issue. – Environmental Pollution 17: 157-167.
- [7] Cao, T., Wang, S., Chen, B. (2019): Water shortage risk transferred through interprovincial trade in Northeast China. – Energy Procedia 158: 3865-3871.
- [8] Causse, J., Billen, G., Garnier, J., Henri-des-Tureaux, T., Olasa, X., Thammahacksa, C., Latsachak, K. O., Soulileuth, B., Sengtaheuanghoung, O., Rochelle-Newall, E., Ribolzi, O. (2015): Field and modelling studies of Escherichia coli loads in tropical streams of montane agro-ecosystems. – Journal of Hydro Environment Research 9: 496-507.
- [9] CCME (2012): Canadian Water Quality Guidelines for the Protection of Aquatic Life and Sediment Quality Guidelines for the Protection of Aquatic Life. – Canadian Council of Ministers of the Environment, Toronto.
- [10] Cheng, N., Liu, L., Hou, Z., Wu, J., Wang, Q., Fu, Y. (2021): Pollution characteristics and risk assessment of surface sediments in the urban lakes. – Environmental Science and Pollution Research 28: 22022-22037.
- [11] Cloern, J. E., Foster, S. Q., Kleckner, A. E. (2014): Phytoplankton primary production in the world's estuarine-coastal ecosystems. – Biogeosciences 11: 2477-2501.
- [12] Dabrowski, J. M., De Klerk, L. P. (2013): An assessment of the impact of different land use activities on water quality in the upper Olifants River catchment. – Water SA. 39(2): 231-244.
- [13] Deng, C., Liu, L., Li, H., Peng, D., Wu, Y., Xia, H., Zhang, A., Zhu, Q. (2021): A datadriven framework for spatiotemporal characteristics, complexity dynamics, and environmental risk evaluation of river water quality. – Science of the Total Environment 785: 147134.
- [14] Dorleku, M. K., Affum, A. O., Tay, C. K., Nukpezah, D. (2019): Assessment of nutrients levels in groundwater within the lower Pra Basin of Ghana. – Ghana Journal of Science 60: 24-36.
- [15] Edokpayi, J. N., Odiyo, J. O., Msagati, T. A. M., Potgieter, N. (2015): Temporal variations in physico-chemical and microbiological characteristics of Mvudi River, South Africa. – International Journal of Environmental Research and Public Health 12(4): 4128- 4140.
- [16] Fernandes, A. C. P., Martins, L. M. O., Pacheco, F. A. L., Fernandes, L. F. S. (2021): The consequences for stream water quality of long-term changes in landscape patterns: implications for land use management and policies. – Land Use Policy 109: 105679.
- [17] Garzon-Vidueira, R., Rial-Otero, R., Garcia-Nocelo, M. L., Rivas-Gonzalez, E., Moure-Gonzalez, D., Fompedrina-Roca, D., Vadillo-Santos, I., Simal-Gandara, J. (2020): Identification of nitrates origin in Limia River basin and pollution-determinant factors. – Agriculture, Ecosystem and Environment 290: 106775., Y., Liu, Z., García-Gir
- [18] Ge on, J., Chen, X., Yan, Y., Li, Z., Xie, Z. (2022): Human-induced loss of functional and phylogenetic diversity is mediated by concomitant deterministic processes in subtropical aquatic insect communities. – Ecological Indicators 136: 108600.
- [19] Gqomfa, B., Maphanga, T., Shale, K. (2022): The impact of informal settlement on water quality of Diep River in Dunoon. – Sustainable Water Resources Management 8: 27.
- [20] Gradilla-Hernández, M. S., Anda, J., Garcia-Gonzalez, A., Meza-Rodríguez, D., Montes, C. Y., Perfecto-Avalos, Y. (2020): Multivariate water quality analysis of Lake Cajititlán, Mexico. – Environmental Monitoring and Assessment 192: 5.
- [21] Griffin, N. J. (2017): The rise and fall of dissolved phosphate in South African rivers. South African Journal of Science 113: (11/12).
- [22] Harding, W. R. (2015): Living with eutrophication in South Africa: a review of realities and challenges. – Transactions of the Royal Society of South Africa 70(2): 155-171.

http://www.aloki.hu ● ISSN 1589 1623 (Print) ● ISSN 1785 0037 (Online)

DOI: http://dx.doi.org/10.15666/aeer/2203_24112423

© 2024, ALÖKI Kft., Budapest, Hungary

- [23] Islam, M. M. M., Islam, M. A. (2020): Quantifying public health risks from exposure to waterborne pathogens during river bathing as a basis for reduction of disease burden. – Journal of Water and Health 18(3): 292-305.
- [24] Jiang, J. P., Sharma, A., Sivakumar, B., Wang, P. (2014): A global assessment of climatewater quality relationships in large rivers: an elasticity perspective. – Science of the Total Environment 468: 877-891.
- [25] Kaushal, S. S., Likens, G. E., Pace, M. L., Utz, R. M., Haq, S., Gorman, J., Grese, M. (2018): Freshwater salinization syndrome on a continental scale. – Proceedings of the National Academy of Sciences 115(4): 574-583.
- [26] Kumar, M., Puri, A. (2012): A review of permissible limits of drinking water. Indian Journal of Occupational Environmental Medicine 16(1): 40-44.
- [27] Laurent, J., Mazumder, A. (2014): Influence of seasonal and inter-annual hydrometeorological variability under varying land-use composition. – Water Research 48: 170-178.
- [28] Liu, Z., Zhou, T., Cui, Y., Li, Z., Wang, W., Chen, Y., Xie, Z. (2021): Environmental filtering and spatial processes equally contributed to macroinvertebrate metacommunity dynamics in the highly urbanized river networks in Shenzhen, South China. – Ecological Process 10(1): 23.
- [29] Mello, K., Valente, R. A., Randhir, T. O., Vettorazzi, C. A. (2018): Impacts of tropical forest cover on water quality in agricultural watersheds in southeastern Brazil. – Ecological Indicators 93: 1293-1301.
- [30] Nartey, V. K., Hayford, E. K., Ametsi, S. K. (2012): Assessment of the impact of solid waste dumpsites on some surface water systems in the Accra Metropolitan Area, Ghana. – Journal of Water Resources Protection 4: 605.
- [31] Nguyen, T. T. N., Némery, J., Gratiot, N., Strady, E., Tran, V. Q., Nguyen, A. T., Aimé, J., Peyne, A. (2019): Nutrient dynamics and eutrophication assessment in the tropical river system of Saigon-Dongnai (southern Vietnam). – Science of The Total Environment 653: 370-383.
- [32] Nyamangara, J., Jeke, N., Rurinda, J. (2013): Long-term nitrate and phosphate loading of river water in the Upper Manyame Catchment, Zimbabwe. – Water SA 39(5): 637-642.
- [33] Olive, G. (2021): Water analysis of the Poncia stream (Gembloux, Belgium). Academia Letters, Article 1952: 1-8.
- [34] Orem, W., Newman, S., Osborne, T. Z., Reddy, K. R. (2015): Projecting changes in Everglades soil biogeochemistry for carbon and other key elements, to possible 2060 climate and hydrologic scenarios. – Environmental Management 55(4): 776-798.
- [35] Romanelli, A., Soto, D. X., Matiatos, I., Martínez, D. E., Esquius, S. (2020): A biological and nitrate isotopic assessment framework to understand eutrophication in aquatic ecosystems. – Science of the Total Environment 715: 136909.
- [36] SANS (South Africa National Standard) (2015): Drinking Water. Part 1: Microbiological, Physical, Aesthetic and Chemical Determinands. Part 2: Application of SANS 241-1. – South African Bureau of Standards (SABS), Pretoria.
- [37] SAWS (South African Weather Service) 2023. Annual State of the Climate of South Africa 2022. – SAWS, Pretoria.
- [38] Singkran, N. (2017): Determining overall water quality related to anthropogenic influences across freshwater systems of Thailand. – International Journal of Water Resources Development 33: 132-151.
- [39] Stimie, C., Richters, E., Thompson, H., Perret, S., Matete, M., Abdallah, K., Kau, J., Mulibana, E. (2001): Hydro-institutional mapping in the Steelpoort river basin, South Africa. Working Paper 17 (South Africa Working Paper No. 6). – International Water Management Institute, Colombo, Sri Lanka.
- [40] Trinh, D. A., Luu, T. N. M., Trinh, Q. H., Tran, H. S., Tran, T. M., Le, T. P. Q., Duong, T. T., Orange, D. (2015): Impact of terrestrial runoff on organic matter, trophic state, and phytoplankton in a tropical, upland reservoir. – Aquatic Sciences 78: 367-379.
- [41] Van Ginkel, C. E. (2011): Eutrophication: present reality and future challenges for South Africa. – Water SA 37(5): 693-701.
- [42] Varol, M., Ustaoğlu, F., Tokatlı, C. (2022): Ecological risks and controlling factors of trace elements in sediments of dam lakes in the Black Sea Region (Turkey). – Environmental Research 205: 112478.
- [43] WHO (World Health Organization) (2011): Guidelines for Drinking-Water Quality. 4th Ed. – WHO, Geneva.
- [44] Zhou, Y., Xu, J. F., Yin, W., Ai, L., Fang, N. F., Tan, W. F., Yan, F. L., Shi, Z. H. (2017): Hydrological and environmental controls of the stream nitrate concentration and flux in a small agricultural watershed. – Journal of Hydrology 545: 355-366.